

De: [Carmen Espinosa](#)
A: [Lorenzo Proia](#)
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Date: mar, 8 dic 2020 a las 17:08
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To: Carmen Espinosa Angona <c.e.angona@gmail.com>

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Title: Effects of the interaction between nutrient concentration and DIN:SRP ratio on geosmin production by freshwater biofilms
Journal: Science of the Total Environment

Dear Miss Carmen Espinosa Angona,

I am pleased to inform you that your paper "Effects of the interaction between nutrient concentration and DIN:SRP ratio on geosmin production by freshwater biofilms" has been accepted for publication in STOTEN and forwarded to the publishers.

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Sincerely,

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Science of the Total Environment

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Science of the Total Environment

Effects of the interaction between nutrient concentration and DIN:SRP ratio on geosmin production by freshwater biofilms

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Abstract:	<p>The global increase of cyanobacterial blooms occurrence has been associated with the presence of compounds that generate earthy and musty odour in freshwater systems, among which geosmin stands out. The lack of information on the factors associated to geosmin production by benthic organisms has driven the development of this study, whose main goal is to determine the effects of nutrient concentration and DIN:SRP ratio on geosmin formation and release.</p> <p>The experiment was performed in 18 microcosms under controlled conditions for 21 days, using a natural biofilm suspension from Ter river (NE, Spain) to promote biofilm settlement. Six treatments were set crossing three DIN:SRP ratios (A = 4:1, B = 16:1 and C = 64:1) with two nutrient concentrations (Low and High).</p> <p>After 7 days of experiment, geosmin was detected in biofilm, being higher under high nutrient concentration and low DIN:SRP ratio conditions. In this treatment, geosmin in biofilm reached its maximum concentration at day 16 (3.8 ± 0.9 ng/mg), decreasing at the end of the experiment (21d) due to cyanobacteria detachment and geosmin release into the water (136 ± 6 ng/L).</p> <p>Overall, this experimental study showed that high nutrient concentration and low DIN:SRP ratio favoured the <i>Oscillatoria</i> genus development within biofilm communities, generating the optimal conditions for geosmin production. The interaction between these two factors was demonstrated to be a potential driver of benthic geosmin production and release, and should be monitored and controlled in rivers exploited for drinking water purposes.</p>
Response to Reviewers:	<p>Reviewer #1: I note and welcome the changes you have made in response to the comments from myself and the other reviewer.</p> <p>The figures are greatly improved from the original submission. Much clearer and more easily interpretable.</p> <p>However, I note there remain a number of English and grammar errors. It appears that there has been an oversight in proof reading since the majority of the revisions have been made and I suggest that the whole article is reviewed thoroughly for such errors. Whilst I acknowledge that a major revision has already been made to this manuscript, I consider that further minor corrections are still necessary prior to the acceptance of this ac</p> <p>Thank you very much for the positive feedback. We have proceeded to correct the</p>

1 **Effects of the interaction between nutrient concentration and**

2 **DIN:SRP ratio on geosmin production by freshwater biofilms**

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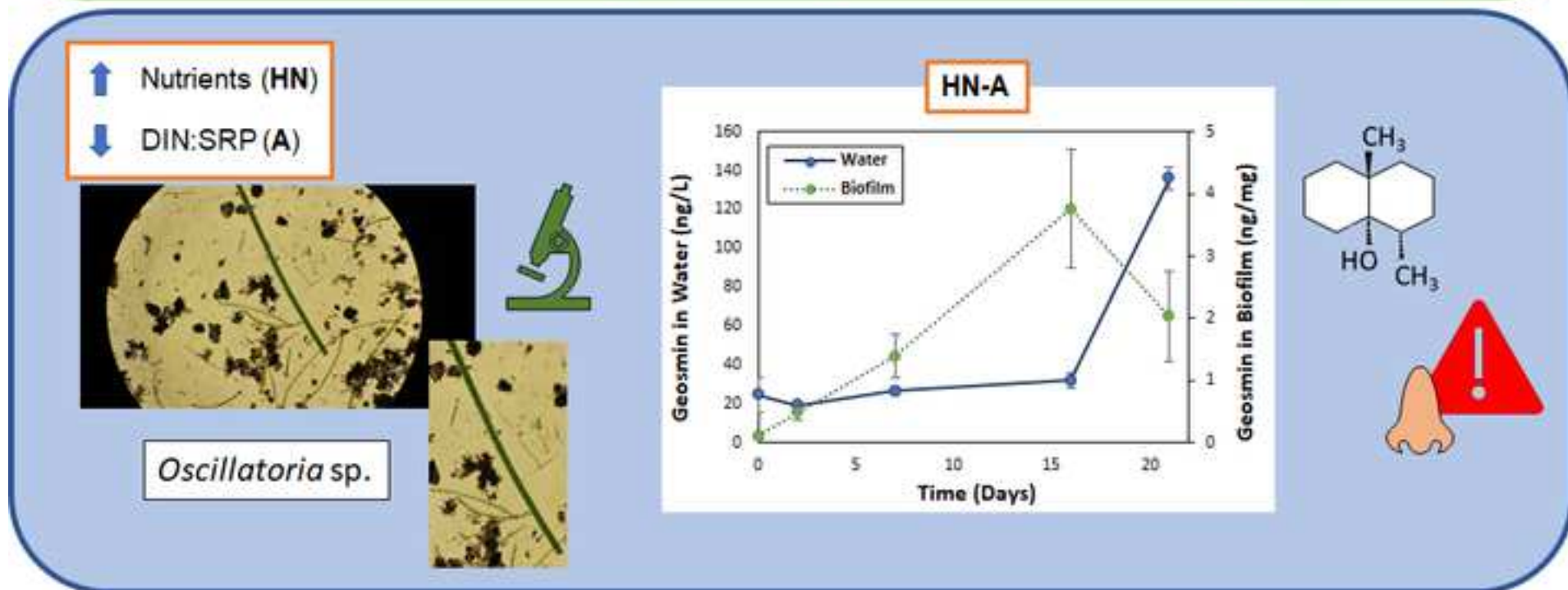
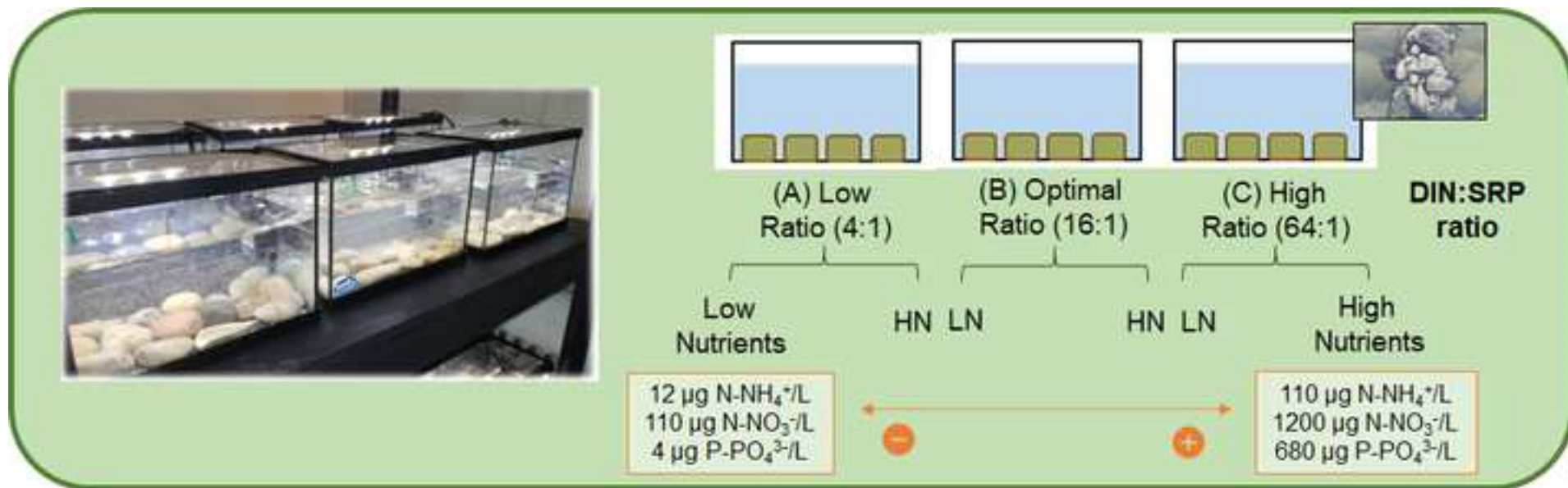
Abstract:

The global increase of cyanobacterial blooms occurrence has been associated with the presence of compounds that generate earthy and musty odour in freshwater systems, among which geosmin stands out. The lack of information on the factors associated to geosmin production by benthic organisms has driven the development of this study, whose main goal is to determine the effects of nutrient concentration and DIN:SRP ratio on geosmin formation and release.

The experiment was performed in 18 microcosms under controlled conditions for 21 days, using a natural biofilm suspension from Ter river (NE, Spain) to promote biofilm settlement. Six treatments were set crossing three DIN:SRP ratios (A = 4:1, B = 16:1 and C = 64:1) with two nutrient concentrations (Low and High).

After 7 days of experiment, geosmin was detected in biofilm, being higher under high nutrient concentration and low DIN:SRP ratio conditions. In this treatment, geosmin in biofilm reached its maximum concentration at day 16 (3.8 ± 0.9 ng/mg), decreasing at the end of the experiment (21d) due to cyanobacteria detachment and geosmin release into the water (136 ± 6 ng/L).

Overall, this experimental study showed that high nutrient concentration and low DIN:SRP ratio favoured the *Oscillatoria* genus development within biofilm communities, generating the optimal conditions for geosmin production. The interaction between these two factors was demonstrated to be a potential driver of benthic geosmin production and release, and should be monitored and controlled in rivers exploited for drinking water purposes.



Highlights

- Geosmin episodes are associated with benthic and planktonic cyanobacterial blooms.
- Low DIN:SRP ratio and high nutrient concentration favor *Oscillatoria* sp development.
- *Oscillatoria* sp produces geosmin under low DIN:SRP ratio and high nutrient conditions.
- Phosphorus concentration is a key factor in geosmin production by benthic cyanobacteria.
- Changes in DIN:SRP due to climate change and human pressures favor geosmin production potential

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32 be monitored and controlled in rivers exploited for drinking water purposes.

33 **Keywords:** Geosmin; *Oscillatoria* sp.; DIN:SRP ratio; nutrient concentration; biofilm;
34 microcosms.

35 **1. Introduction**

36 In recent decades, anthropogenic activities have severely degraded streams and rivers
37 worldwide. Excessive nutrient load is considered one of the major threats leading to
38 substantial alterations in the overall functioning and structure of the ecosystem (Doods and
39 Smith, 2016). These effects are of greater importance in rivers affected by situations of water
40 scarcity (i.e. Mediterranean streams), where water flow can be extremely reduced, thus
41 generating less dilution capacity and a greater increase in nutrient concentration (Karaouzas et
42 al., 2018). The variations in nutrient export induced by anthropogenic activities can deeply
43 modify the ratios between dissolved inorganic nitrogen (DIN) and soluble reactive phosphorus
44 (SRP) (von Schiller et al., 2008), with the consequent exposure of freshwater microbial
45 communities to significant fluctuations of both nutrients' concentration and balance over time
46 (Artigas et al., 2015). These modifications of nutrients availability and balance can provoke a

47 marked increase of algal communities' growth in aquatic environments, leading to
48 cyanobacterial blooms typical of eutrophicated ecosystems.

49 Abundant growth of benthic and planktonic cyanobacteria in freshwaters exploited for
50 drinking water purposes has been correlated with the appearance of secondary metabolites,
51 such as geosmin and 2 – methylisoborneol (MIB) (Espinosa et al., 2020; Lee et al., 2017),
52 affecting the organoleptic characteristics of water and influencing both odor and taste, with a
53 consequent negative impact in the perception of the population about the tap water quality
54 (Ding et al., 2014). Although no health risk has been linked to these Taste and Odor
55 compounds (T&Os), they are perceived by humans at very low levels of concentration (Smith
56 et al., 2009), resulting in consumer complaints and representing a challenge for drinking water
57 companies. Among the existing T&Os, geosmin has been identified as the main metabolite
58 leading to bad taste and odor in drinking waters (Watson et al., 2016). This metabolite is a
59 volatile bicyclic terpenoid, produced by certain cyanobacterial cells during the exponential
60 phase of growth and released into the water as a consequence of cell death and/or biomass
61 decomposition (Kim et al., 2018; Lee et al., 2017). Although some heterotrophic bacteria have
62 been identified as potential geosmin producers (such as the actinomycete bacteria
63 *Streptomyces* sp. and the fungi *Penicillium* sp.), they usually have a terrestrial origin or are
64 present in sediments and foodstuffs. And while they can also be related to some geosmin
65 episodes in aquatic systems, cyanobacteria are still considered the main geosmin producers in
66 freshwater ecosystems (Lukassen et al., 2019; Olsen et al., 2016). The majority of these
67 geosmin-producing cyanobacteria are benthic or epiphytic (70%), while the rest are planktonic
68 (Jüttner and Watson, 2007). In the group of geosmin-producing cyanobacteria, *Oscillatoria* sp.,
69 *Lynghya* sp., *Symploca* sp. and *Dolichospermum* sp. have been identified as the most common
70 genera in freshwater ecosystems (Smith et al., 2009). These genera have also been described
71 as MIB producers, although only a few species have the ability to produce both geosmin and
72 MIB (Jüttner and Watson, 2007). Even if numerous episodes have been reported all over the

73 world, the main triggers driving the production of these compounds by cyanobacteria in
74 freshwaters are not deeply understood yet (Tung et al., 2008).

75 Fluvial biofilms are complex microbial communities formed by different groups of organisms,
76 which together with extracellular enzymes and detritus are enclosed within a polymeric matrix
77 (Romaní, 2010). The effects of nutrient concentration on the biofilm community of running
78 waters have been widely documented describing that, depending on the nutrient
79 concentrations, the stoichiometry of the benthic biofilm can differ, conjointly with the nutrient
80 uptake rates (consequence of changes in nutrient demand) modifying the microbial
81 interactions between the autotrophic and heterotrophic biofilm compartments (Price and
82 Carrick, 2016; Bechtold et al., 2012). Furthermore, deviation from Redfield ratio (106C:16N:1P)
83 has been used as an indication of nutrient limitation for algal growth (Redfield et al., 1963).
84 Specifically, the N:P ratio of 16:1 is used as a reference to differentiate P-limitation (N:P>16)
85 from N-limitation (N:P<16) although this value may differ between algal and cyanobacterial
86 groups (Geider and La Roche, 2002; Sabater et al., 2016). Therefore, water N:P ratios may
87 control population dynamics and the coexistence of species in river biofilms.

88 Previous studies indicated that environmental factors such as temperature, light and nutrient
89 availability influence geosmin production (Alghanmi et al., 2018; Lee et al., 2017). However,
90 most of these were carried out on single species cultures under highly controlled conditions
91 (Parinet et al., 2010; Li et al., 2012; Suurnäkki et al., 2015). On the other hands, some field
92 studies have suggested that high nutrient concentrations, low nitrogen to phosphorus ratio
93 (N:P) and relatively low NO₃:NH₃ ratio can be important factors potentially promoting algal
94 growth, and favoring the cyanobacterial production of secondary metabolites such as geosmin
95 and MIB (Olsen et al., 2016; Harris et al., 2016). Most of these studies have focused on
96 phytoplanktonic organisms, whereas very few have deal with the identification of geosmin
97 production drivers in natural river benthic communities. A field study performed in the

98 Llobregat river (NE Spain) pointed out that the nutrient imbalance (TN:TP = 10) could have
99 favored geosmin production by biofilms (Vilalta et al., 2003). However, field studies alone are
100 not enough to establish causal relationships between environmental factors and biological
101 responses and need to be confirmed under controlled conditions. Therefore, there is still a lack
102 of information about the independent and combined effects of the availability and imbalance
103 of nutrients on triggering geosmin production by benthic microorganisms growing within
104 biofilm communities.

105 Nowadays, the water quality of more and more rivers is deteriorating due to the increase in
106 the nutrients' concentration derived from anthropogenic activities. As has been described in
107 reservoirs, this nutrient increase can lead to the uncontrolled growth of cyanobacteria, which
108 in shallow rivers develop mainly as benthic. These rivers can be a source of water for drinking
109 water treatment companies, so the drivers behind cyanobacteria growth, such as nutrient
110 concentrations and DIN:SRP ratio, need to be evaluated in order to generate useful
111 information to understand and predict the occurrence of geosmin episodes in these
112 freshwater ecosystems.

113 This study was designed to explore the independent and combined effects of the DIN:SRP ratio
114 and nutrient concentrations on the biofilm structure-function, and its relationship with the
115 geosmin production and release. To this end, an experiment was carried out in controlled
116 laboratory microcosms exposing natural biofilm communities to three different DIN:SRP ratios
117 - low (4:1), optimal (Redfield, 16:1) and high (64:1) - under two nutrient concentration levels
118 (low and high) in water. Although some studies have been carried out evaluating geosmin
119 production by different cyanobacteria strains in laboratory cultures, to the best of our
120 knowledge this is the first study evaluating the independent and combined effect of the
121 DIN:SRP ratio and nutrient concentration on geosmin production by a natural benthic biofilm.
122 The main outcomes of this study could be very useful for catchment managers and water

123 utilities, helping them to understand under which nutrient conditions they should expect
124 geosmin in water used for drinking purposes, improving their response time in geosmin
125 treatment and reducing consumer discomfort and complaints.

126 **2. Materials and Methods**

127 **2.1. Experimental design and sampling**

128 Eighteen microcosms consisting of 6L glass aquariums (length x width x height 26 x 15 x 17 cm)
129 were used to evaluate the biofilm functional and structural response under different nutrient
130 concentrations and DIN:SRP ratios in water. Each microcosm, containing 16 scraped and
131 autoclaved stream cobbles, was filled with 3L of artificial water. Artificial water was prepared
132 to simulate a pristine stream as described in Ylla et al. (2009) and was obtained by dissolving
133 pure salts in distilled water creating a chemical composition of 14.96 mg/L Na⁺, 10.81 mg/L
134 Ca²⁺, 0.52 mg/L K⁺, 0.40 mg/L Mg²⁺, 9.71 mg/L SO₄²⁻, 12.45 mg/L SiO₃²⁻, 19.67 mg/L Cl²⁻ and
135 14.52 mg/L HCO₃⁻. To maintain these conditions, the water of the microcosms was renewed
136 every two to three days. Each aquarium had a submersible pump (EDEN 105, Eden Water
137 Paradise, Italy) to promote oxygenation and water circulation. The photoperiod was set at 12h
138 light: 12h dark using LEDs (LENB 135-lm, LENB/14.97/11.98), and the room temperature was
139 set at 16°C.

140 The experiment was carried out during a geosmin episode that occurred in March 2019 in the
141 Ter river (NE Spain), coinciding with the presence of benthic *Oscillatoria* mats (authors,
142 personal observation). From this river, a natural biofilm suspension was obtained by scraping
143 several cobbles randomly selected along a 50m river stretch located at the water collection
144 point of the water drinking company Aigües de Vic (X440478; Y4648779 UTM31N), and 15 mL
145 of this suspension was inoculated in each microcosm to promote biofilm colonization. During
146 the colonization period (three weeks), a new inoculum was added after each water renewal to
147 favor biofilm settlement in the microcosms (Vendrell-Puigmitja et al., 2020).

148 After the colonization period, the exposure period started. Six treatments (n = 3) were set by
149 crossing three DIN:SRP ratios (Low = 4:1 (A), Optimal = 16:1 (B) and High = 64:1 (C)) with two
150 nutrient concentrations (Low and High) (**Table 1**). Specifically, these treatments were low
151 nutrient concentration and low DIN:SRP ratio (LN-A), low nutrient concentration and optimal
152 DIN:SRP ratio (LN-B), low nutrient concentration and high DIN:SRP ratio (LN-C), high nutrient
153 concentration and low DIN:SRP ratio (HN-A), high nutrient concentration and optimal DIN:SRP
154 ratio (HN-B) and high nutrient concentration and high DIN:SRP ratio (HN-C). DIN:SRP ratios
155 were determined as DIN (Dissolved Inorganic Nitrogen) divided by Soluble Reactive
156 Phosphorus (SRP) in molar quantities. DIN concentration was determined as the sum of
157 ammonium (N-NH₄⁺) and nitrate (N-NO₃⁻) concentration. Nutrient concentration and DIN:SRP
158 ratios were chosen to cover a range representative of Ter river basin variation, whose nutrient
159 concentration ranges between 12 – 300 µg N-NH₄⁺/L, 100 – 2000 µg N-NO₃⁻/L, and 10 – 950 µg
160 P-PO₄³⁻/L (authors, field study; Water Catalan Agency, 2020). During the exposure period,
161 which lasted for 21 days, nutrient concentration in each aquarium was maintained at each
162 water renewal, which were done every two to three days. Nitrogen concentration was
163 adjusted using nitrate and ammonium standard solutions (1 g/L PanReac-AppliChem), and the
164 phosphorus concentration was adjusted using a phosphate concentrated stock solution (10
165 mM P-PO₄³⁻) obtained by dissolving pure salts (KH₂PO₄, PanReac-AppliChem) into deionized
166 water.

167 Before and after each water renewal during both the colonization and exposure periods,
168 physico-chemical parameters were measured with portable probes: temperature, dissolved
169 oxygen concentration and saturation (YSI professional plus, YSI Incorporated, USA), pH (XS
170 pH7+ DHS) and electrical conductivity (XS COND 7+). From each microcosm, water samples
171 were taken and filtered through 0.2 µm nylon membrane filters (Merck Millipore) before the
172 analysis of SRP and DIN forms (ammonium and nitrate). All samples were stored at -20°C until
173 analysis.

174 Biofilms were sampled just before the beginning of the exposure period (t₀) and after 2, 7, 16
175 and 21 days. Three cobbles were randomly sampled from each microcosm at each sampling
176 day. The photosynthetic efficiency and the phototrophic community composition were
177 measured directly with an amplitude modulated fluorometer (Mini-PAM fluorometer, Walz,
178 Effeltrich, Germany) and a BenthosTorch portable fluorometer probe (bbe Moldaenke,
179 Schwentineta, DK) respectively. After that, each cobble was scrapped in 45 mL of water from
180 the same microcosm to obtain a biofilm suspension. Aliquots of this suspension were used to
181 analyze geosmin concentration in biofilm, Chlorophyll *a* (Chl *a*), ash free dry mass (AFDM),
182 algal taxonomic composition and extracellular enzymatic activities (AEE). AEE samples were
183 analyzed the same day and the rest were stored frozen (-20°C) until analyses, except aliquots
184 for geosmin concentration, which were stored at -80°C and taxonomic samples, which were
185 fixed with formalin (2%) and stored at 4°C. The area of the scraped cobbles was obtained by
186 drawing the surface on aluminum foil and recalculating depending on the weight (Graham et
187 al., 1988).

188 Water samples were also taken at each sampling time to analyze nutrients and geosmin
189 concentrations in water. Additionally, on the last day of the experiment, a nutrient uptake
190 capacity experiment was performed with the colonized cobbles present in the microcosms.
191 Samples for geosmin quantification in water were stored at 4°C in dark conditions until the
192 analysis, which was performed within 48 hours after collection to avoid degradation. Water
193 samples for nutrient analysis were filtered through 0.2 µm nylon membrane filters (Merck
194 Millipore) and frozen at -20°C until analysis.

195 **2.2. Biofilm samples**

196 **2.2.1. Geosmin**

197 The concentration of geosmin in biofilm was quantified following the protocol described by
198 Espinosa et al. (2020). Briefly, a headspace solid phase micro-extraction was performed. To

199 obtain 50 mL of total volume, 5 mL of the biofilm sample were decanted in a 100 mL opaque
200 reaction vial containing (i) 45 mL of saline solution (10 g of NaCl in 45 mL of sterile dH₂O) and
201 (ii) a magnetic stir bar. The samples were frozen (-80°C) and thawed five times to facilitate cell
202 breakage and geosmin release in the aquatic phase. A 65 µm PDMS/DVB fiber was injected
203 into the headspace of the sample vial, and the vial was placed on a magnetic stirrer inside an
204 oven at 60°C for 25 min to extract geosmin from the samples. Geosmin characterization was
205 performed using a GC/MS instrument (ISQ – TRACE GC ULTRA). Geosmin was separated using a
206 capillary column (Sigma Aldrich SPB®-5 Capillary GC Column) measuring 30 m x 0.25 mm with a
207 film thickness of 0.25 µm. Carrier gas was helium at a flow rate of 1 mL/min. Geosmin was
208 desorbed by exposing the fiber in the GC injection port for 6 min at 250°C. A 0.75 mm internal
209 diameter glass liner was used, and the injection port was in splitless mode. The temperature
210 program was isothermal for 3 min at 70°C, raised to 200°C at a rate of 10°C/min and, finally, to
211 280°C at a rate of 30°C/min, with a hold time of 4 min. The transfer line to the mass
212 spectrometer and the ion source was 250 and 200°C respectively, the scan was in SIM mode,
213 with the scanned fragments being 111, 112, 125, 164 and 182 m/z. The analytical detection
214 limit was 2.5 ng/L, and the precision of the method was evaluated with the relative standard
215 deviation (RSD ≤20%). Geosmin concentration in biofilm was expressed as ng geosmin/mg
216 AFDM.

217 **2.2.2. Chlorophyll *a***

218 Algal biomass in biofilms was estimated from the analysis of Chlorophyll *a* (Chl *a*)
219 concentration. Chlorophyll *a* was extracted from the sample with 90% acetone over a period of
220 12 h in the dark at 4°C. Acetone extracts were filtered through 0.7 µm glass fiber filters (GF/F
221 filters, Whatman International) and Chlorophyll *a* concentration was determined
222 spectrophotometrically measuring the absorbance at 430, 665 and 750nm (NanoPhotometer™)

223 P-360, IMPLEN) following the method described by Jeffrey & Humphrey (1975). Chlorophyll *a*
224 concentration was expressed as $\mu\text{g}/\text{cm}^2$.

225 **2.2.3. AFDM**

226 Total biofilm biomass was measured as ash free dry mass (AFDM). The aliquots of biofilm
227 suspension were filtered through pre-combusted (4 h at 500°C, Carbolite muffle ELF 11/14B)
228 and pre-weighed filters (47-mm GF/F Whatman glass-fiber filters, 0.7 μm pore size), then dried
229 for 72 h at 60°C (Forced air oven, MEMMERT IFE500) in order to calculate dry mass (DM).
230 Afterwards, samples were combusted at 500°C (Carbolite muffle ELF 11/14B) for 4 h, and then
231 weighed again. The differences in filter mass before and after drying were calculated for DM
232 (72 h at 60°C) and after combustion (4 h at 500°C) subtracted from DM to obtain AFDM.
233 Results were standardized by the scratched biofilm surface of the cobbles and expressed as
234 g/m^2 .

235 **2.2.4. Algal taxonomic composition in biofilm**

236 Algal identification was done following Lange Bertalot (2001), counting a minimum of 400 cells
237 and measuring at least 6 vision fields per sample. Cell biovolume was obtained following the
238 procedure described in Hillebrand et al. (1999). To calculate the biovolume, the length and
239 width of a minimum of ten individuals, randomly selected, were measured for each genus. The
240 optical microscope (Nikon Eclipse 600W) using phase-contrast and Nomarski differential
241 interference contrast optics at a magnification of 400 increments was used, counting and
242 measuring the cells of at least 6 fields of vision per sample. The values were transformed as
243 described in Ricart et al. (2009), using the surface of the sample, the dilutions needed, and the
244 number of cells counted, and were expressed as $\mu\text{m}^3/\text{cm}^2$.

245 **2.2.5. Extracellular Enzymatic Activities**

246 Leucine-aminopeptidase (EC 3.4.11.1) and phosphatase (EC 3.1.3.1–2) activity was measured
247 by using fluorescent-linked substrates (L-leucine-7-amino-4-methylcoumarin hydrochloride
248 [AMC] used for peptidase and 4-methylumbellyferyl-phosphate [MUF] for phosphatase)
249 following the protocol described in Romaní & Marxsen (2002). The fluorescence was measured
250 at 380/460 nm excitation/emission for MUF and at 380/460 nm excitation/emission for AMC
251 using the Spark® multimode microplate reader (TECAN). Values are expressed as μmol
252 $\text{MUF}/\text{cm}^2\cdot\text{h}$ and $\mu\text{mol AMC}/\text{cm}^2\cdot\text{h}$.

253 **2.3. Water samples**

254 **2.3.1. Geosmin**

255 To analyze geosmin concentration in water samples, 50 ml of each sample and 10 g of NaCl
256 were added to a 100 mL opaque reaction vial. To extract the geosmin in water a 65 μm
257 PDMS/DVB fiber was used. Separation and analysis of the extracted volatile compound was
258 performed in a GC/MS instrument (ISQ – TRACE GC ULTRA), as described in section 2.2.1.
259 Geosmin concentration in water was expressed as ng/L.

260 **2.3.2. Nutrients**

261 SRP concentration in water samples was measured using the protocol described by Murphy &
262 Riley (1962). DIN concentration was determined as the sum of ammonium (N-NH_4^+) and nitrate
263 (N-NO_3^-) concentration. N-NH_4^+ was analyzed following the protocol described in Reardon et al.
264 (1966), and N-NO_3^- was determined following the method described in Rand et al. (1976). For
265 all these colorimetric determinations, the absorbance was measured with the
266 spectrophotometer NanoPhotometer™ P-360, INPLEM.

267 **2.3.3. Nutrient Uptake**

268 The nutrient uptake capacity of biofilms was calculated by measuring ammonium, nitrate and
 269 phosphate temporal decay after a controlled spike in microcosms. The artificial water of each
 270 microcosm was renewed to ensure that the starting conditions of the aquariums were optimal.
 271 Each microcosm was then spiked with an appropriate volume of a phosphate concentrated
 272 stock solution (10 mM P-PO₄³⁻) obtained by dissolving pure salts (KH₂PO₄, PanReac-AppliChem)
 273 into deionized water, and ammonium (NH₄⁺, 1 g/L standard solution, PanReac-AppliChem)
 274 stock solution in order to quadruple the basal concentration of nutrients (Proia et al., 2017).
 275 The biofilms were then incubated for 4 hours under controlled temperature and light
 276 conditions. Aliquots of water were taken at 1, 5, 30, 60, 120, 180 and 240 minutes after
 277 spiking, immediately filtered through 0.2 µm nylon membrane filters (Merk Millipore) and
 278 stored at -20°C until analysis. The uptake capacity (U, 1/m²·min) was calculated as (Eq. 1):

$$279 \quad U = \frac{[(C_i \cdot V_i) - (C_f \cdot V_f)]}{\frac{C_i \cdot V_i}{A \cdot T_f}} \quad (\text{Eq. 1})$$

280 where C_i = initial nutrient concentration; V_i = initial volume; C_f = final nutrient concentration; V_f
 281 = final volume; A = biofilm's area and T_f = residence time.

282 **2.4. Data treatment**

283 Before the statistical analysis, the Kolmogorov – Smirnov test was performed to verify that the
 284 variables fulfilled the conditions of normal distribution; and if they did not, they were
 285 logarithmically transformed. Physicochemical and biological data were evaluated using the
 286 two-way repeated measures analysis of variance (ANOVA) in SPSS Statistics version 21, with
 287 nutrient concentration and DIN:SRP ratio as factors and sampling date (time) as repeated
 288 measure. In addition, physicochemical and biological data was evaluated on each sampling
 289 date by a two-way ANOVA, with the DIN:SRP ratio and nutrient concentration as independent
 290 variables, and geosmin in water (ng/L), geosmin in biofilm (ng/mg), cyanobacteria and diatoms

291 biomass ($\mu\text{g}/\text{cm}^2$), Chlorophyll *a* ($\mu\text{g}/\text{cm}^2$), photosynthetic efficiency (Y_{eff}) and PHO (μmol
292 MUF/ $\text{cm}^2\cdot\text{h}$) as dependent variables. The post-hoc test performed was the Bonferroni test.
293 Pearson correlation coefficient tests were carried out to explore the relationship between the
294 variables. Statistical significance was set at $p < 0.05$ for all tests performed. The distribution of
295 the variables according to the treatments was evaluated using a principal component analysis
296 (PCA) performed with RStudio software (version 3.6.0).

297 **3. Results**

298 **3.1. Physicochemical parameters**

299 The physicochemical conditions in the microcosms remained stable throughout the
300 experiment, with water temperature varying insignificantly around 16°C, slight alkaline pH and
301 low electrical conductivity (**Table 1**), and without significant differences among them despite
302 the treatments (repeated measures ANOVA).

303 **3.2. Geosmin concentration**

304 **3.2.1. Geosmin in biofilm**

305 Geosmin concentration in biofilm (**Figure 1. A.**, **Table 2**) varied significantly over time among
306 microcosms, with the HN-A treatment being the one which significantly differed from the
307 others (**Table 3**). From day 7 to the end of the experiment, geosmin concentration in biofilm
308 was significantly the highest in the HN-A treatment (**Table 2**). Under these conditions, geosmin
309 concentration increased gradually, reaching its maximum (3.8 ± 0.9 ng/mg) at $t=16\text{d}$ and
310 decreasing thereafter to 2.0 ± 0.7 ng/mg ($t=21\text{d}$, **Figure 1.A.**). Under HN-B conditions, the
311 biofilm did not produce geosmin during the whole experiment.

312 **3.2.1. Geosmin in water**

313 At the beginning of the exposure period (t=0d), geosmin was detected in all treatments at low
314 concentrations (27.4 ± 6.7 ng/L), without any significant difference among microcosms (**Figure**
315 **1.B.**, **Table 2**, **Table 3**). After 7 days of treatment, geosmin concentration in water was
316 significantly higher in those treatments with a lower DIN:SRP ratio (Bonferroni test: $p < 0.005$)
317 independently of the nutrient concentration. However, from then on, the interaction between
318 the two factors had a significant effect on geosmin concentration. At the end of the
319 experiment (day 21), both nutrient concentration and the DIN:SRP ratio, as well as their
320 interaction showed a significant effect on geosmin concentration in water. Specifically, the
321 microcosms developed under the lowest DIN:SRP conditions showed higher geosmin
322 concentration in water compared to the other ratios (Bonferroni test: $p < 0.001$), and the
323 highest geosmin concentration (136 ± 6 ng/L) was found under HN conditions (**Figure 1.B.**).
324 The only treatment where geosmin was not detected at the end of the experiment was HN-B.

325 **3.3. Biofilm attributes**

326 **3.3.1. Structural parameters**

327 **3.3.1.1. Community composition**

328 During the experiment, the phototrophic community composition of the biofilm
329 (cyanobacteria, green algae and diatoms) varied over time (**Table 2**). Specifically, the DIN:SRP
330 ratio significantly influenced the cyanobacteria biomass and the relative diatoms and
331 cyanobacteria abundances over time, while the diatoms biomass was affected by the nutrient
332 concentration and its interaction with the DIN:SRP ratio (**Table 3**).

333 One week after the beginning of the exposure period (t=7d), the cyanobacteria biomass was
334 statistically different between treatments depending on the DIN:SRP ratio, being higher under
335 low DIN:SRP treatments (Bonferroni test: $p < 0.01$). At day 16, nutrient concentration and its
336 interaction with the DIN:SRP ratio also had a significant effect on cyanobacteria biomass, being

337 significantly higher under low DIN:SRP treatments (Bonferroni test: $p < 0.01$) and resulting the
338 highest in HN-A treatment ($1.63 \pm 0.36 \mu\text{g chl}a/\text{cm}^2$). A similar trend was observed for the
339 cyanobacteria relative abundance. The highest value of abundance was found in the HN-A
340 treatment at $t=16\text{d}$ ($48 \pm 2\%$), although at the end of the experiment, this value decreased to
341 $31 \pm 7\%$ (**Figure 2**). The opposite pattern was observed for the diatom's relative abundance.
342 At the end of the experiment ($t=21\text{d}$) the algal taxonomic community in the biofilm was
343 evaluated (**Figure 3**). The relative abundance of *Oscillatoria* sp. was higher under low DIN:SRP
344 ratio conditions compared with the other ratios (Bonferroni test: $p = 0.001$), reaching the
345 highest value ($45 \pm 10\%$) under the HN-A treatment. The cyanobacterium *Oscillatoria* sp. and
346 the diatom *Melosira* sp. biovolume varied among microcosms depending on the DIN:SRP ratio
347 ($F = 18.9$; $p < 0.001$ for *Oscillatoria* sp., $F = 17.5$; $p < 0.001$ for *Melosira* sp.) and its interaction
348 with nutrient concentration ($F = 5.0$; $p < 0.05$ for *Oscillatoria* sp., $F = 5.0$; $p < 0.05$ for *Melosira*
349 sp.). In addition, a strong negative correlation was found between *Oscillatoria* sp. and *Melosira*
350 sp. biovolume (Pearson's correlation: $r = -0.914$; $p < 0.001$). The *Oscillatoria* sp. biovolume was
351 also positively correlated with geosmin concentration in water and biofilm (Pearson's
352 correlation: $r = 0.886$; $p < 0.001$ and $r = 0.888$; $p < 0.001$, respectively).

353 **3.3.1.2. Chlorophyll a**

354 Chlorophyll a concentration varied significantly throughout the experiment depending on the
355 treatment (**Table 2**, **Table 3**). At the end of the experiment ($t=21\text{d}$), both factors and their
356 interaction showed a significant effect on the biofilm Chlorophyll a content (*Supplementary*
357 *Figure*), with higher concentration under the optimal DIN:SRP ratio conditions compared with
358 high DIN:SRP ratio treatment (Bonferroni test: $p < 0.005$). It was also observed that under HN
359 conditions the biofilm had a higher production of Chlorophyll a , with the HN-B treatment being
360 the one showing the highest concentration ($18.3 \pm 0.9 \mu\text{g}/\text{cm}^2$) (**Table 2**).

361 **3.3.2. Functional parameters**

362 **3.3.2.1 Photosynthetic capacity**

363 The photosynthetic capacity of the biofilm (expressed as Y_{eff}) (**Table 2**) was affected by the
364 DIN:SRP ratio, showing at $t=16d$ significant lower values under low DIN:SRP treatments (0.563
365 ± 0.025) compared to the optimal DIN:SRP ratio (0.646 ± 0.016) (Bonferroni test: $p < 0.05$)
366 (**Table 3**). Five days later, nutrient concentration also affected the Y_{eff} values, with the lowest
367 being found for the biofilms developed under high nutrient concentrations and low DIN:SRP
368 ratio conditions (0.529 ± 0.046).

369 **3.3.2.2. Nutrient uptake**

370 Phosphate uptake (or release) capacity was significantly affected by nutrient concentration (F
371 $= 5.4$, $p < 0.05$) and its interaction with DIN:SRP ratio ($F = 25.4$, $p < 0.001$). In general, the
372 uptake capacity was higher in LN conditions ($0.20 \pm 0.06 \text{ mg/m}^2\cdot\text{min}$) than under HN conditions
373 ($0.09 \pm 0.04 \text{ m}^2\cdot\text{min}^{-1}$), except for the high DIN:SRP ratio that showed the opposite trend. The
374 ammonium uptake capacity was significantly affected by nutrient concentration ($F = 25.2$, $p <$
375 0.001), with higher uptake capacities under LN conditions ($0.37 \pm 0.07 \text{ m}^2\cdot\text{min}^{-1}$) compared
376 with HN ($0.13 \pm 0.004 \text{ m}^2\cdot\text{min}^{-1}$). Nitrate uptake capacity did not differ among microcosms.

377 **3.3.2.3. Extracellular Enzymatic Activities**

378 The phosphatase activity (PHO) of the biofilms varied throughout the experiment (**Table 3**).
379 Specifically, at $t=16d$, the biofilms under high DIN:SRP ratio conditions showed higher PHO
380 than those of the optimal DIN:SRP ratio treatment (Bonferroni test: $p < 0.05$) (**Table 2**). At the
381 end of the experiment, the PHO values were different depending on the nutrient
382 concentration and DIN:SRP ratio, with the high DIN:SRP ratio treatment being significantly
383 different from the others (Bonferroni test: $p < 0.05$), and higher under HN conditions ($72.0 \pm$
384 $18.4 \text{ } \mu\text{mol MUF/cm}^2\cdot\text{h}$). For the leucine-aminopeptidase activity, no significant differences
385 were observed among microcosms over time.

386 Both nutrients uptake/release rates and enzymatic activities showed differences between
387 treatments depending on the nutrient concentration under which the biofilm developed,
388 without any relationship with the production and release of geosmin.

389 **3.3.3. Overall relationships between biofilm responses and geosmin dynamics**

390 The Principal Component Analysis (**Figure 4**) shows how the parameters evaluated are
391 distributed according to the different treatments. It emphasizes that the X axis (explaining the
392 49.8% of variance) separates the HN-A treatment from the rest, being the one with the highest
393 concentration of geosmin in water and in biofilm, greater cyanobacteria biomass, and higher
394 presence of *Oscillatoria* sp. The Y axis (17.8% of variance) separates the optimal (B) and high
395 (C) DIN:SRP ratio treatments according to the concentration of nutrients (LN or HN), since
396 these conditions affected the enzymatic activities and the nutrients uptake/release capacities
397 of the biofilms.

398 **4. Discussion**

399 This study demonstrates that, both nutrient concentration and DIN:SRP ratio are important
400 drivers of geosmin occurrence in freshwaters, since they favor the growth of geosmin-
401 producing organisms in biofilm communities and promoting intracellular geosmin formation
402 and its subsequent release to the water column.

403 The biofilm community structure began to change 7 days after the nutrient manipulation
404 started, mainly because of the different DIN:SRP ratios. In particular, cyanobacteria increased
405 its relative abundance, becoming predominant in the biofilms exposed to low DIN:SRP ratio
406 conditions, especially under high nutrient concentration (HN-A) (**Figure 2**). These results agree
407 with recent studies describing how the phosphorus availability excess may stimulate the
408 magnitude of cyanobacteria blooms (Jankowiak et al., 2019). Moreover, it has been described
409 that certain cyanobacteria can store essential metabolic nutrients in the cytoplasm, especially

410 under eutrophic conditions (Felisberto et al., 2011). For example, phosphorus can be
411 accumulated in intracellular polyphosphate granules, whereas under nitrogen-limited
412 conditions some cyanobacteria have also shown a high capacity to fix it (Felisberto et al.,
413 2011). Different studies, performed mainly in lakes, have pointed out some nutrient thresholds
414 to control the growth of cyanobacteria, such as the TP value, which had to be between 20 –
415 100 µg TP/L (Li et al., 2018; Sharma et al., 2011). Stroom and Kardinaal (2016) reported that
416 below 0.03 mg TP/L, the risk of cyanobacteria dominance is 10%, increasing to 40% at 0.07 mg
417 TP/L, whereas another study indicated that N levels can be also important, 0.8 mg TN/L being
418 the threshold to limit the growth rate of *Microcystis* dominated blooms (Xu et al., 2014). The
419 results obtained in this study point out that similar values can be fixed as nutrient thresholds
420 for cyanobacteria growth in rivers, together with the N:P ratio, since the N:P ratio value should
421 be maintained at a level of at least 10, but preferably above 50 – 64, thereby reducing the
422 likelihood for N-limiting situation which may favor cyanobacteria dominating blooms (Li et al.,
423 2018). It should be noted that in our study we have evaluated the N:P ratio as DIN:SRP, while
424 in the studies of lakes and reservoirs the ratio is TN:TP.

425 At the end of the experiment (t=21d), cyanobacteria abundance in the HN-A treatment
426 decreased, arguably because of cell degradation. It has been demonstrated that the growth
427 phase of different cyanobacterial strains could last from 8 to 24 days before entering the
428 stationary or degradation phase (Kruskopf & Du Plessis, 2006; Jindal et al., 2011).
429 Nevertheless, the conditions were still optimal for the cyanobacteria community, which were
430 present in this treatment, with *Oscillatoria* sp. accounting for almost 45% of the total biofilm
431 biovolume, a rather high value compared to its presence in other treatments (**Figure 3**). The
432 growth phase of the *Oscillatoria* genus varies depending on the species. Kruskopf & Du Plessis
433 (2006) observed that *Oscillatoria simplicissima* reached the fast growth phase after 8 days,
434 whereas *Oscillatoria formosa* could grow exponentially for 24 days and before starting the
435 stationary phase (Jindal et al., 2011).

436 Changes at the structural level were reflected in functional parameters with photosynthetic
437 efficiency (Y_{eff}) being significantly lower under low DIN:SRP ratio. This could be explained by
438 the higher presence of cyanobacteria in these treatments, since some studies have found that
439 cyanobacteria usually show low photosynthetic efficiency values compared with the rest of the
440 algal community (Espinosa et al., 2020; Luimstra et al., 2018).

441 In this study, the highest geosmin concentration in both biofilm and water was registered
442 under low DIN:SRP ratio combined with high nutrient concentration (HN-A), thus confirming
443 the important role of these variables as major drivers of the biofilm geosmin production and
444 release. Moreover, our results corroborate that cyanobacteria, and more specifically
445 *Oscillatoria* sp., are responsible for geosmin production in freshwater biofilm communities. In
446 fact, the linear regression analyses performed on cyanobacteria biomass ($\mu\text{g chl}a/\text{cm}^2$) and
447 geosmin occurrence in biofilms and water show a significantly high relationship (**Figure 5**). In
448 our study, geosmin concentration pattern differed between the two compartments evaluated
449 (water and biofilm) over time (**Figure 1**). In HN-A treatment, geosmin in biofilm increased
450 earlier than in water (**Figure 1.A.**), and reached its maximum at $t = 16\text{d}$ resulting in $99.1 \pm 0.6\%$
451 of the total geosmin detected (both water and geosmin). This was the maximum geosmin
452 concentration observed in biofilm and coincided with the greater cyanobacteria abundance in
453 the biofilm (**Figure 2**). In the same treatment, the geosmin concentration in biofilm decreased
454 substantially at the end of the experiment, accounting for $86.6 \pm 5.1\%$ of total geosmin
455 detected. This was reflected in a marked increase of dissolved geosmin concentration in water
456 (**Figure 1.B.**). The decoupling observed between the presence of geosmin in biofilm and its
457 release to the water could explain the lower R^2 value obtained in the linear regression analysis
458 between cyanobacteria biomass and geosmin in water. This pattern could be related with the
459 cyanobacteria life cycle, and more specifically to their optimal geosmin production (associated
460 with the growth phase) and release time (linked to biomass decomposition and/or cell lysis),
461 since it has been described that many cyanobacteria species can produce geosmin, although

462 they may not actively release it until cell lysis occurs (Kim et al., 2018; Sabater et al., 2014).

463 Furthermore, the relative amounts of intra and extracellular portions of geosmin may also vary

464 considerably with cell age, environmental conditions and among different species (Alghanmi et

465 al., 2018). The results obtained in our study agreed with other studies carried out with

466 cyanobacterial cultures, which indicated that the highest production of geosmin occurred in

467 the late exponential growth phase, with release starting during the stationary phase and full

468 release occurring with cell's death and lysis (Saadoun et al., 2001). Another study performed

469 with *Anabaena* sp. cultures found that about 85 – 95% of the total geosmin was concentrated

470 in the cell rather than released in the cultivation medium (Alghanmi et al. 2018). A similar

471 trend was observed by Ho et al. (2012), who found that >98% of the total geosmin was

472 intracellular in untreated *Anabaena circinalis* rich waters. Nevertheless, all these studies were

473 carried out with monospecific cyanobacterial cultures, whereas our work is the first one

474 demonstrating the role of nutrient concentration and DIN:SRP ratio as the main drivers of the

475 geosmin intracellular production and release to the water from producing organisms that are

476 part of a complex benthic microbial community.

477 At the structural level, Chlorophyll *a* concentration was affected by both factors and its

478 interaction, with higher Chlorophyll *a* concentration under HN-B conditions. There was a

479 negative correlation between geosmin and Chlorophyll *a* concentration in the biofilm

480 (Pearson's correlation: $r = -0.535$; $p = 0.049$). This trend could be observed at the end of the

481 experiment ($t=21d$), when the highest concentration of Chlorophyll *a* (0.38 ± 0.01 ng/mg) was

482 measured in the treatment in which there was no geosmin production (HN-B) (**Figure 6**).

483 Geosmin and Chlorophyll *a* have the same metabolic pathway and it has been described that

484 when cyanobacteria start synthesizing geosmin, the production of Chlorophyll *a* decreases or

485 even stops (Cai et al., 2017). Other studies have shown that under elevated nitrate

486 concentrations, a greater amount of Chlorophyll *a* is synthesized, with a corresponding

487 decrease of geosmin synthesis (Saadoun et al., 2001). This partially agrees with our results,

488 since under elevated nitrogen conditions, we found the highest Chlorophyll *a* concentration.
489 However, when phosphorus concentration also increased (lower DIN:SRP ratio), Chlorophyll *a*
490 concentration was lower and geosmin levels increased (**Figure 6**). This evidence would confirm
491 the fundamental role of phosphorus concentration in the geosmin production process.

492 The results of this study indicate that the interaction between the DIN:SRP ratio and nutrient
493 concentration in water has a key role in the geosmin production and its release by
494 cyanobacteria in the biofilm. Therefore, freshwater systems affected by high nutrient
495 concentration and with an imbalance of phosphorus over nitrogen are of special concern
496 because they are susceptible to experience geosmin episodes. These episodes could lead to
497 bad odor and taste situations on surface waters, potentially affecting the customers trust and
498 being a huge problem for water utilities in those rivers exploited for drinking purposes.

499 **5. Conclusions and perspectives**

500 Overall, this experimental study shows that both nutrient concentration and DIN:SRP ratio has
501 a clear effect on the biofilm community structure and function, and consequently on the
502 geosmin formation in the biofilm and its subsequent release into water. Low DIN:SRP ratio and
503 high nutrient concentration favored the appearance of the *Oscillatoria* genus, generating the
504 optimal conditions for geosmin production. However, it should be noted that the conclusions
505 of this study are limited to the nutrient levels evaluated, without knowing whether the
506 increase in cyanobacterial biomass as a function of the DIN:SRP ratio and the nutrients
507 concentration is linear, exponential or follows some other pattern. Therefore, further
508 investigations on this topic should evaluate wider variation of these factors to establish the
509 trend of these behaviors with greater precision.

510 Our results could help drinking water companies in the forecasting and management of
511 geosmin episodes in rivers and shallow reservoirs, where the main producers are benthic
512 cyanobacteria, simply by carrying out the DIN:SRP ratio calculation. Furthermore, monitoring

513 the geosmin concentration in biofilm can lead to a notable increase in the ability to advance to
514 an episode of geosmin in water, since this study has shown that after a few days of detecting
515 high concentrations of geosmin in biofilm, it is released into the water. However, today
516 implementing this analysis in the laboratories of drinking water treatment companies could be
517 expensive and time-consuming. Nevertheless, given the remarkable relationship established
518 between geosmin in biofilm and cyanobacteria, by including the evaluation of biofilm
519 community's development in the regular monitoring, drinking water companies could detect
520 the increase of cyanobacteria abundances which, in a situation of higher nutrient availability
521 and low DIN:SRP ratio, could trigger the appearance of geosmin in water later on.

522 The need to predict and manage these geosmin episodes is especially important in areas with
523 Mediterranean climatic conditions under the current global change context, which is expected
524 to drastically reduce river flows as a consequence of more severe droughts. In this context, our
525 results strongly suggest a potential increase of geosmin episodes, which will be favored by the
526 increase of nutrient concentration, consequence of the decreased dilution capacity during
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685

686 **Table 1.** Mean value and standard deviation (n=30) for the ammonium (N-NH₄⁺ µg/L), nitrate (N-NO₃⁻ µg/L), SRP (P-PO₄³⁻ µg/L), the DIN:SRP ratio, pH, water
 687 temperature (°C), electrical conductivity (EC) (µS/cm), dissolved oxygen (DO) (mg/L), and oxygen saturation (%) for the treatments: low nutrient (LN) and
 688 high nutrient (HN); low ratio (A = 4:1), optimal ratio (B = 16:1) and high ratio (C = 64:1).

	Treatment					
	LN			HN		
	A	B	C	A	B	C
N-NH₄⁺ (µg/L)	12.7 ± 2.7	12.6 ± 0.6	11.7 ± 1.7	110 ± 7	110 ± 6	106 ± 5
N-NO₃⁻ (µg/L)	107 ± 15	108 ± 18	112 ± 5	1175 ± 112	1174 ± 90	1142 ± 24
P-PO₄³⁻ (µg/L)	61.0 ± 4.8	18.2 ± 3.8	4.0 ± 0.6	682 ± 49	164 ± 9	42.0 ± 6.6
DIN:SRP	4.2 ± 0.2	14.9 ± 1.4	69.8 ± 7.2	4.4 ± 0.5	16.5 ± 1.1	64.8 ± 5.2
pH	8.1 ± 0.2	8.1 ± 0.2	8.1 ± 0.2	8.0 ± 0.3	8.0 ± 0.2	8.1 ± 0.2
Temperature (°C)	16.8 ± 0.4	16.9 ± 0.3	16.8 ± 0.3	16.7 ± 0.3	16.8 ± 0.3	16.9 ± 0.3
EC (µS/cm)	164 ± 17	163 ± 14	158 ± 33	166 ± 13	164 ± 14	170 ± 12
DO (mg/L)	8.3 ± 0.5	8.0 ± 0.6	8.2 ± 0.5	8.2 ± 0.5	8.4 ± 0.5	8.0 ± 0.4
Saturation (%)	85 ± 5	84 ± 7	85 ± 5	84 ± 5	86 ± 5	83 ± 4

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694 **Table 2.** Mean value and standard deviation for geosmin in water (ng/L), geosmin in biofilm (ng/mg), cyanobacteria and diatoms biomass ($\mu\text{g}/\text{cm}^2$),
695 Chlorophyll *a* ($\mu\text{g}/\text{cm}^2$), photosynthetic efficiency (Yeff) and PHO ($\mu\text{mol MUF}/\text{cm}^2\cdot\text{h}$) at the beginning of the experiment (not significant differences between
696 treatments), and for the treatments: low nutrient (LN) and high nutrient (HN); low ratio (A = 4:1), optimal ratio (B = 16:1) and high ratio (C = 64:1), at times
697 7, 16 and 21 days.

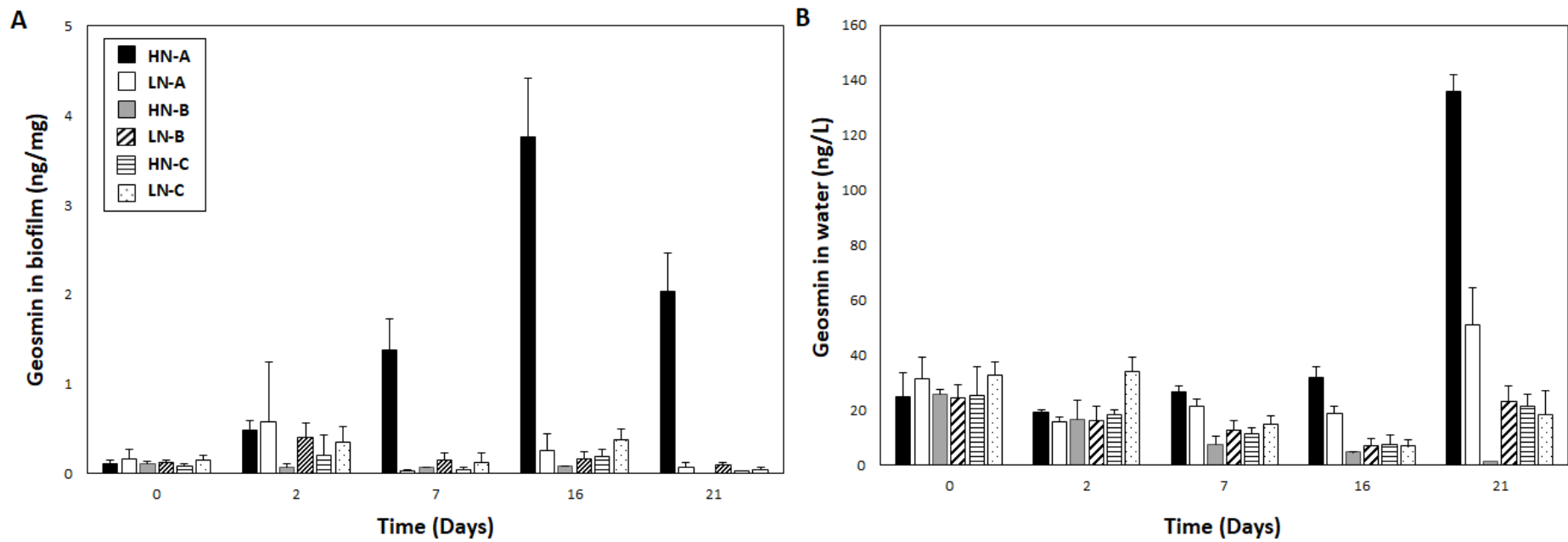
	Treatment																		
	LN									HN									
	A			B			C			A			B			C			
	0d	7d	16d	21d	7d	16d	21d	7d	16d	21d	7d	16d	21d	7d	16d	21d	7d	16d	21d
Geosmin in water (ng/L)	27.4 ± 6.7	21.2 ± 2.9	18.7 ± 2.8	20.9 ± 13.6	12.8 ± 3.4	7.0 ± 2.8	23.3 ± 5.4	14.8 ± 2.9	7.2 ± 2.1	18.4 ± 8.7	26.6 ± 2.2	32.0 ± 3.9	135.8 ± 6.1	7.5 ± 3.1	4.8 ± 0.0	< 2.5	11.6 ± 1.9	7.5 ± 3.5	21.5 ± 4.2
Geosmin in biofilm (ng/mg)	0.10 ± 0.05	0.03 ± 0.01	0.26 ± 0.19	0.08 ± 0.04	0.16 ± 0.08	0.17 ± 0.07	0.10 ± 0.03	0.13 ± 0.1	0.38 ± 0.12	0.04 ± 0.03	1.39 ± 0.35	3.76 ± 0.96	2.03 ± 0.74	0.08 ± 0.00	0.09 ± 0.00	0.00 ± 0.00	0.04 ± 0.03	0.20 ± 0.07	0.03 ± 0.00
Cyanobacteria ($\mu\text{g chl}a/\text{cm}^2$)	0.43 ± 0.06	1.23 ± 0.71	0.61 ± 0.27	0.55 ± 0.06	0.28 ± 0.22	0.38 ± 0.21	0.48 ± 0.20	0.52 ± 0.30	0.52 ± 0.10	0.39 ± 0.20	1.95 ± 0.36	1.63 ± 0.36	0.87 ± 0.38	0.28 ± 0.16	0.57 ± 0.15	0.46 ± 0.05	0.66 ± 0.23	0.34 ± 0.31	0.59 ± 0.14
Diatoms ($\mu\text{g chl}a/\text{cm}^2$)	2.10 ± 0.60	2.68 ± 0.42	1.55 ± 0.82	1.61 ± 0.45	2.37 ± 0.99	1.62 ± 0.63	1.63 ± 0.35	3.00 ± 2.31	2.66 ± 1.29	1.61 ± 0.51	3.49 ± 0.40	2.14 ± 1.12	2.33 ± 1.90	1.65 ± 1.02	4.35 ± 1.05	2.99 ± 2.52	2.00 ± 0.51	1.66 ± 1.69	3.38 ± 0.75
Chlorophyll <i>a</i> ($\mu\text{g}/\text{cm}^2$)	3.90 ± 0.64	3.50 ± 1.31	3.10 ± 1.89	6.51 ± 0.51	2.63 ± 0.48	3.82 ± 1.19	4.20 ± 0.92	7.87 ± 0.88	4.14 ± 1.38	2.80 ± 0.70	2.87 ± 0.81	3.68 ± 0.32	8.68 ± 1.73	5.12 ± 2.45	7.99 ± 1.32	18.32 ± 0.97	5.10 ± 0.28	11.72 ± 2.31	10.00 ± 3.37
Yeff	0.589 ± 0.029	0.646 ± 0.064	0.581 ± 0.083	0.640 ± 0.053	0.614 ± 0.036	0.635 ± 0.068	0.680 ± 0.040	0.639 ± 0.047	0.555 ± 0.056	0.645 ± 0.033	0.630 ± 0.084	0.545 ± 0.042	0.529 ± 0.046	0.607 ± 0.083	0.657 ± 0.025	0.672 ± 0.031	0.633 ± 0.031	0.615 ± 0.036	0.631 ± 0.068
PHO (MUF/$\text{cm}^2\cdot\text{h}$)	19.7 ± 1.9	16.8 ± 7.7	16.5 ± 9.4	24.2 ± 4.2	13.0 ± 1.6	23.6 ± 12.9	29.3 ± 11.6	33.2 ± 8.5	42.0 ± 20.1	31.6 ± 11.7	17.4 ± 9.6	25.6 ± 9.2	35.9 ± 11.7	24.2 ± 8.3	17.5 ± 7.9	36.1 ± 6.5	26.9 ± 14.2	54.4 ± 28.8	72.0 ± 18.4

698

699 **Table 3.** Statistical F and p value for the two-way repeated measures ANOVA and the two-way
700 ANOVA on physicochemical and biological data. Sampling time t=0d and t=2d have not been
701 included since there were not statistical differences for any of the variables evaluated. The
702 factors evaluated were nutrient concentration (N), nutrient ratio (R) and their interaction
703 (NxR) as independent variables, and geosmin in water (ng/L), geosmin in biofilm (ng/mg),
704 cyanobacteria and diatoms biomass ($\mu\text{g chl}a/\text{cm}^2$), Chlorophyll *a* ($\mu\text{g}/\text{cm}^2$), photosynthetic
705 efficiency (Yeff) and PHO ($\mu\text{mol MUF}/\text{cm}^2\cdot\text{h}$) as dependent variables.

		Repeated measures ANOVA		ANOVA					
				t = 7d		t = 16d		t = 21d	
		F	p	F	p	F	p	F	p
Geosmin in Water (ng/L)	N	9.22	<0.001	0.34	n.s.	4.22	n.s.	21.65	0.001
	R	37.03	<0.001	22.54	<0.001	50.42	<0.001	126.28	<0.001
	NxR	15.30	<0.001	3.44	n.s.	7.47	0.01	50.05	<0.001
Geosmin in Biofilm (ng/mg)	N	26.34	<0.001	22.40	0.001	30.63	0.001	6.90	<0.05
	R	22.00;	<0.001	25.14	<0.001	39.07	<0.001	9.64	0.01
	NxR	24.89	<0.001	35.33	<0.001	39.95	<0.001	9.67	0.01
Cyanobacteria ($\mu\text{g chl}a/\text{cm}^2$)	N	2.80	<0.05	0.81	n.s.	5.08	<0.05	0.12	n.s.
	R	6.45	<0.001	5.28	<0.05	33.67	<0.001	3.44	n.s.
	NxR	1.53	n.s.	0.09	n.s.	7.75	<0.01	0.55	n.s.
Diatoms ($\mu\text{g chl}a/\text{cm}^2$)	N	0.98	n.s.	0.93	n.s.	1.75	n.s.	0.73	n.s.
	R	3.16	<0.001	4.39	<0.05	28.45	<0.001	3.44	n.s.
	NxR	1.76	n.s.	0.13	n.s.	12.00	<0.01	1.18	n.s.
Chlorophyll <i>a</i> ($\mu\text{g}/\text{cm}^2$)	N	41.04	<0.001	0.28	n.s.	14.82	<0.01	110.16	<0.001
	R	13.74	<0.001	12.08	<0.01	6.34	<0.05	14.53	0.001
	NxR	13.74	<0.001	6.57	<0.05	3.74	n.s.	22.19	<0.001
Yeff	N	3.72	0.01	0.78	n.s.	4.59	n.s.	2.06	n.s.
	R	4.86	<0.001	0.30	n.s.	6.00	<0.05	5.24	<0.05
	NxR	2.23	<0.05	0.02	n.s.	0.99	n.s.	3.16	n.s.
PHO ($\mu\text{mol MUF}/\text{cm}^2\cdot\text{h}$)	N	1.82	n.s.	0.17	n.s.	0.38	n.s.	11.17	<0.01
	R	0.73	n.s.	3.48	n.s.	4.96	<0.05	5.77	<0.05
	NxR	1.14	n.s.	1.22	n.s.	0.43	n.s.	3.29	n.s.

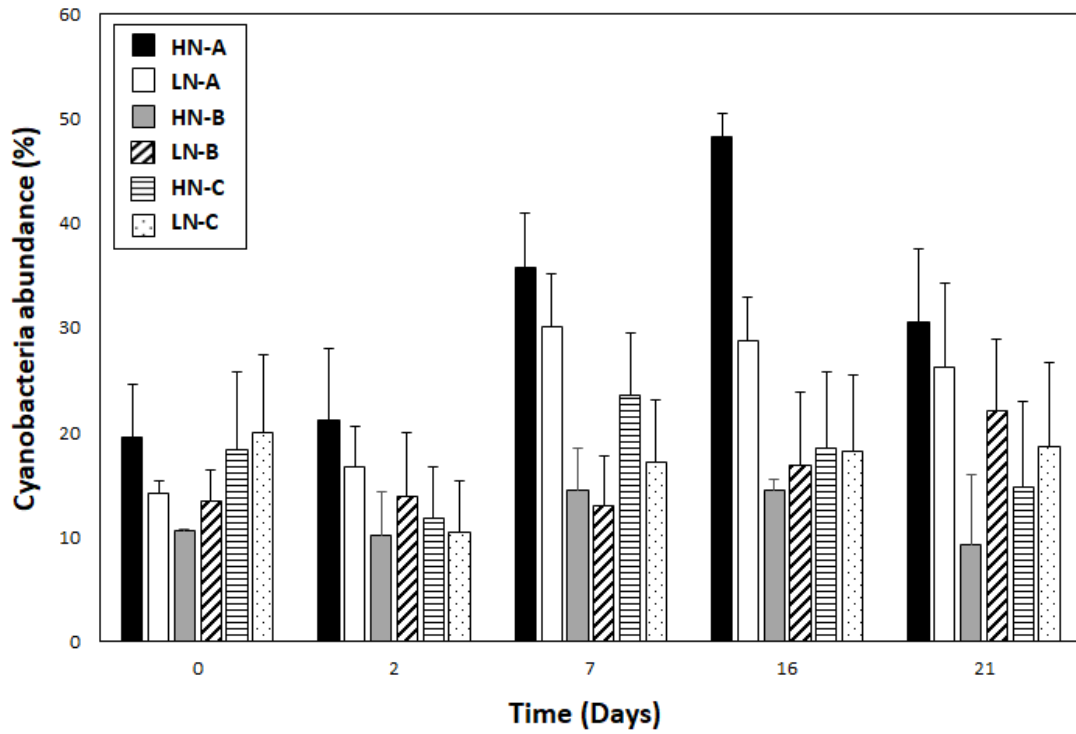
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708 **Figure 1.** Mean values and standard deviation on sampling days t =0, 2, 7 16 and 21 days, for each treatment (HN-A, LN-A, HN-B, LN-B, HN-C and LN-C) for **A.**

709 Geosmin concentration in biofilm (ng/mg) and **B.** Geosmin concentration in water (ng/L).

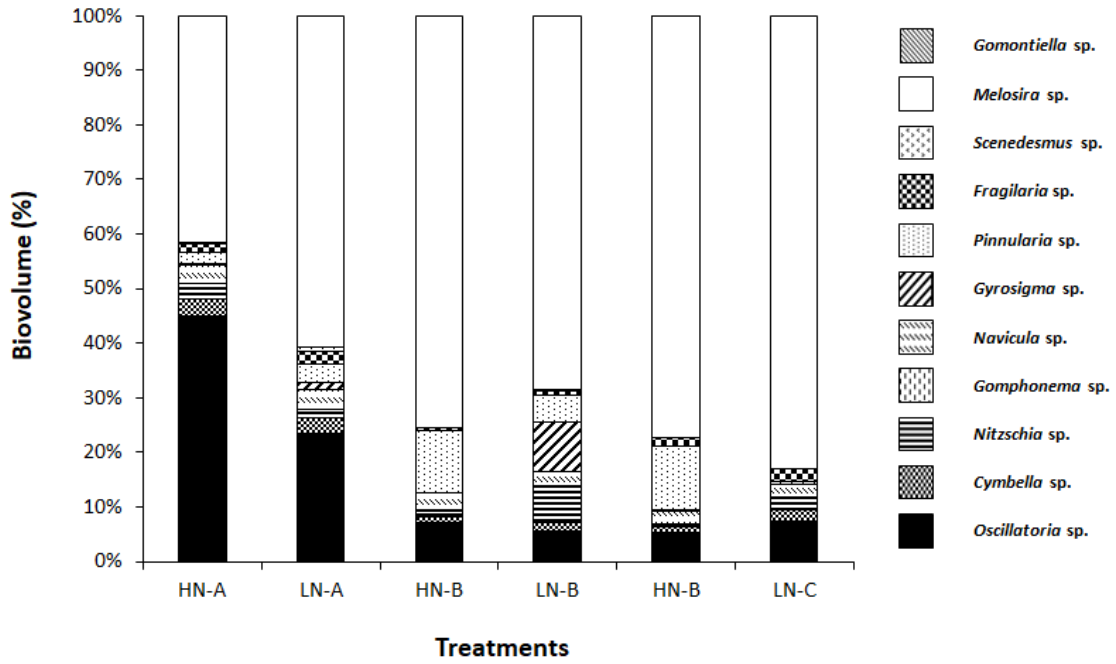


710

711 **Figure 2.** Mean values and standard deviation of the cyanobacteria relative abundance (in %)
 712 given by the BenthosTorch for the sampling days (t = 0, 2, 7 16 and 21 days) for each treatment:
 713 HN-A, LN-A, HN-B, LN-B, HN-C and LN-C.

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717 **Figure 3.** Relative taxa expressed as biovolume (in %) present in the biofilm of each treatment:

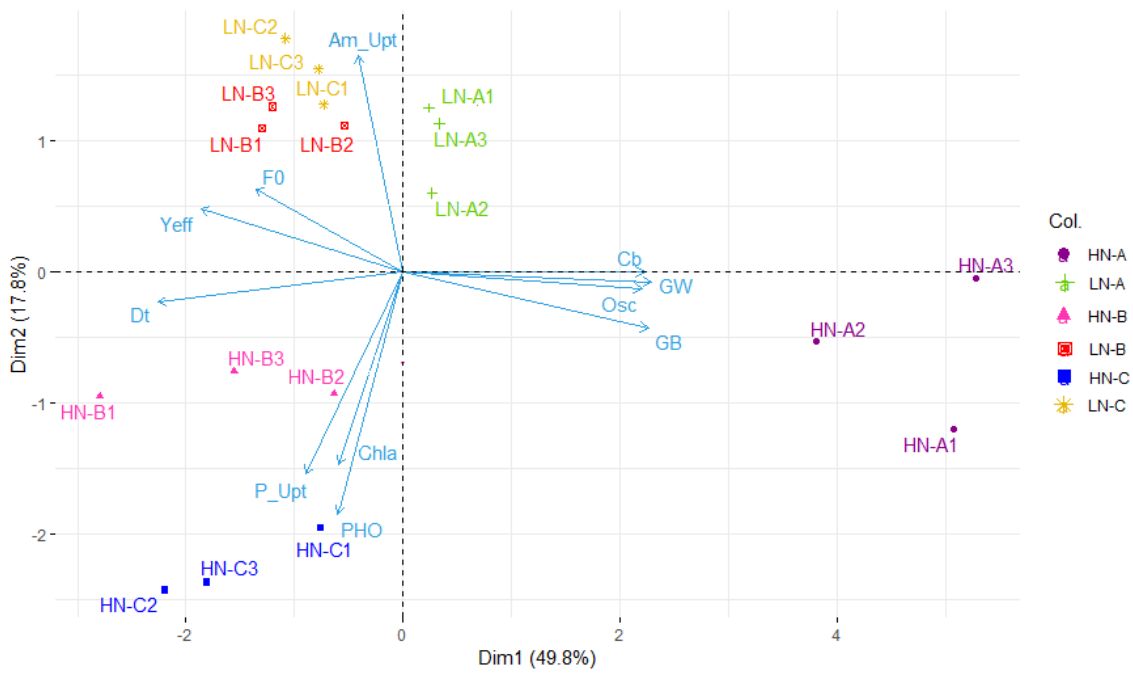
718 HN-A, LN-A, HN-B, LN-B, HN-C and LN-C.

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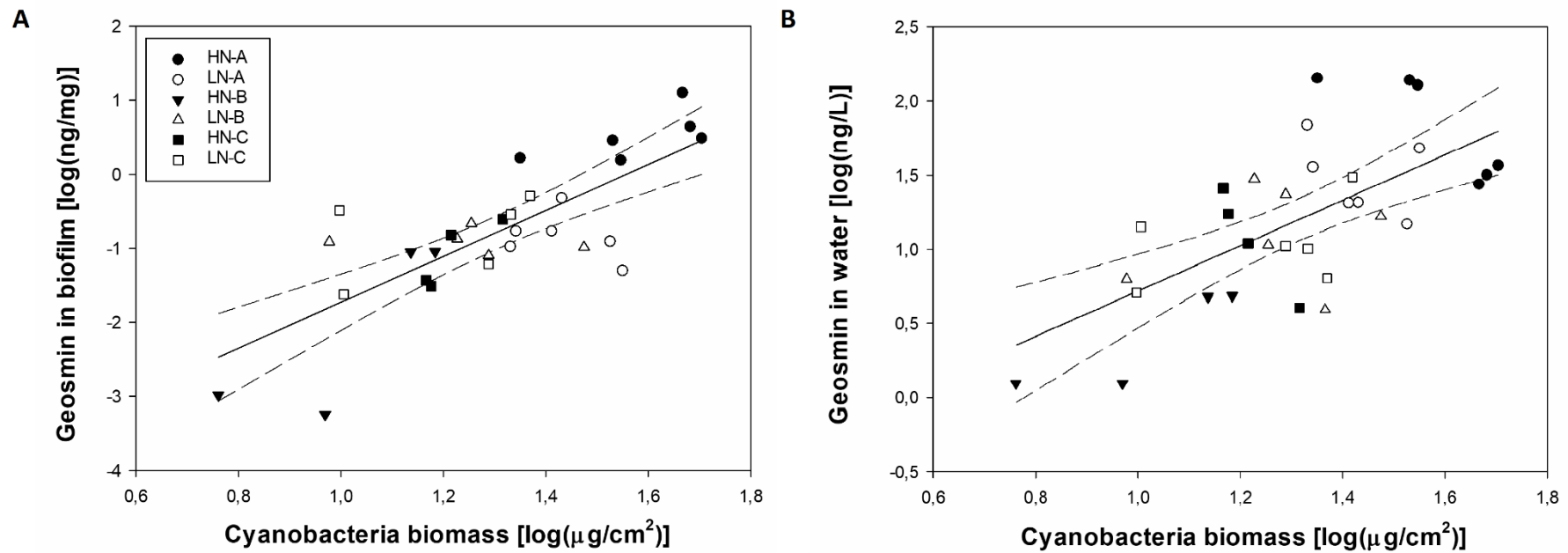
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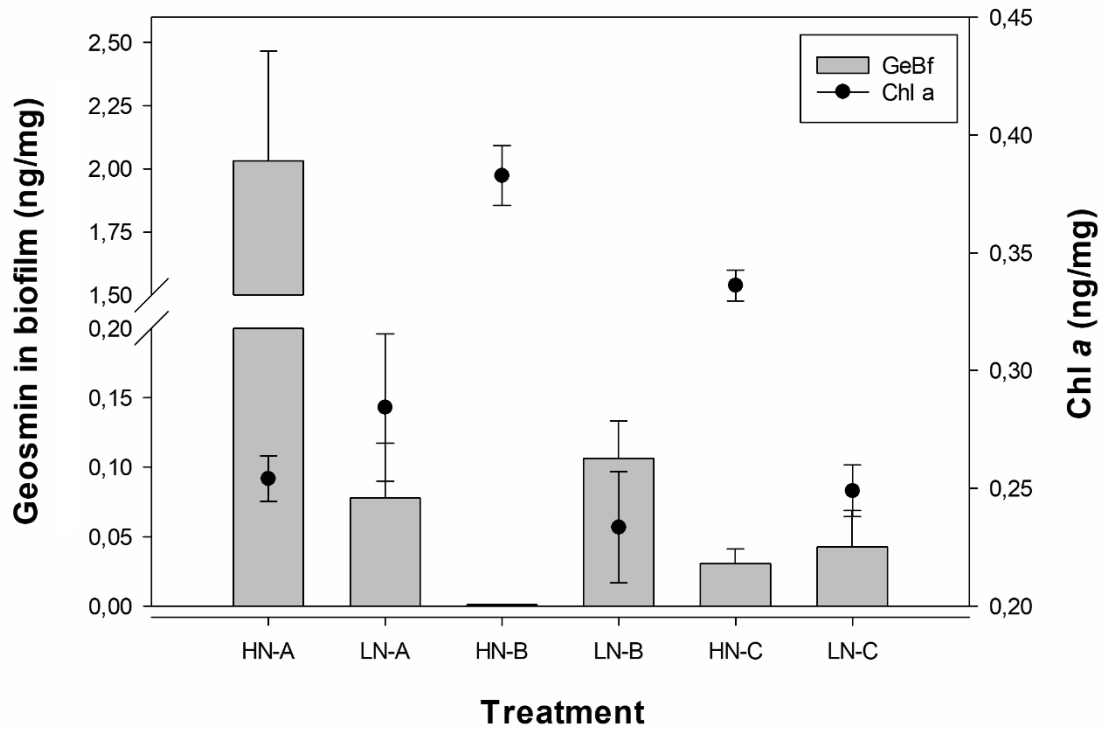
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724 **Figure 4.** Principal Component Analysis showing treatments distribution based on the variables
 725 evaluated. Axes 1 and 2 combined explain 67.6% of the variance. Treatments: HN-A, LN-A, HN-
 726 B, LN-B, HN-C and LN-C. Variables: geosmin in biofilm (GB), geosmin in water (GW),
 727 cyanobacteria biomass (Cb), diatoms biomass (Dt), *Oscillatoria* sp. (Osc), photosynthetic
 728 efficiency (Yeff), Chlorophyll a (Chla), SRP uptake (P-Upt), ammonium uptake (Am_Upt) and
 729 phosphatase activity (PHO).



730

731 **Figure 5.** Relationship between **(A)** log geosmin in biofilm (ng/mg) and log cyanobacteria biomass ($\mu\text{g}/\text{cm}^2$) ($R^2 = 0.59$, $F = 40.35$, $p < 0.0001$) and **(B)** log
 732 geosmin in water (ng/L) and log cyanobacteria biomass ($\mu\text{g}/\text{cm}^2$) ($R^2 = 0.43$, $F = 22.76$, $p < 0.0001$), at $t = 16\text{d}$ and $t = 21\text{d}$.



733

734 **Figure 6.** Mean values and standard deviation at t=21d for each treatment (HN-A, LN-A, HN-B,

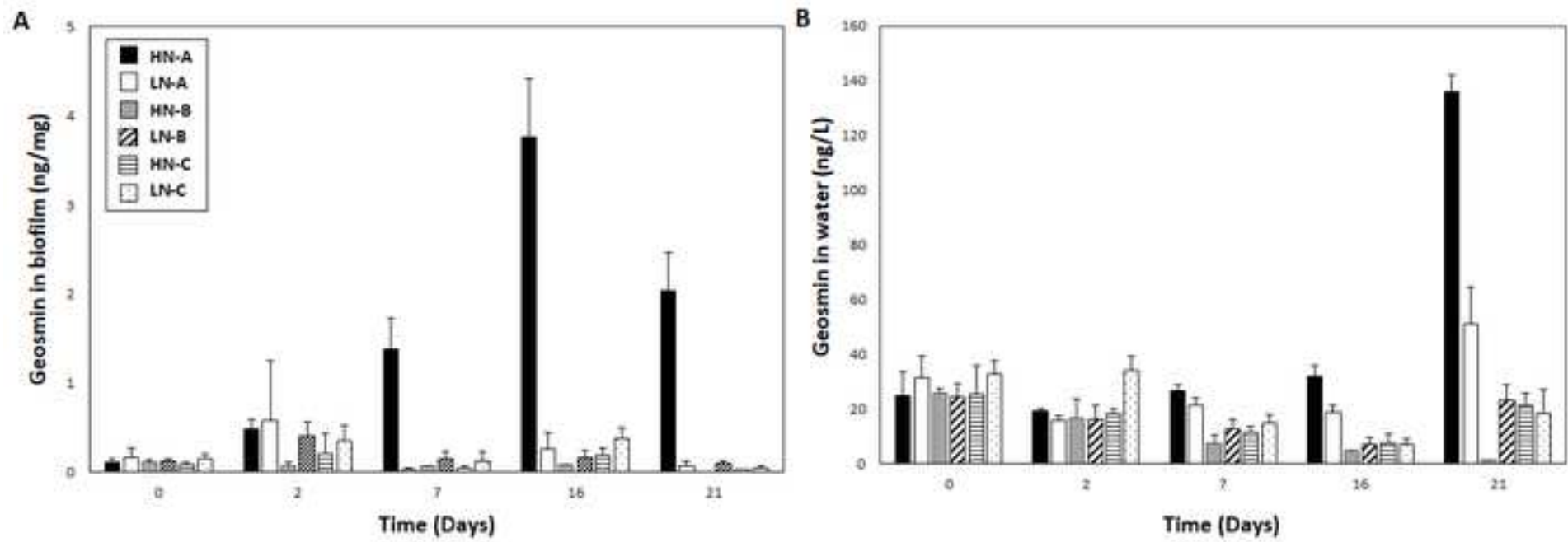
735 LN-B, HN-C and LN-C) of the geosmin concentration in biofilm (ng/mg) vs. Chlorophyll *a*

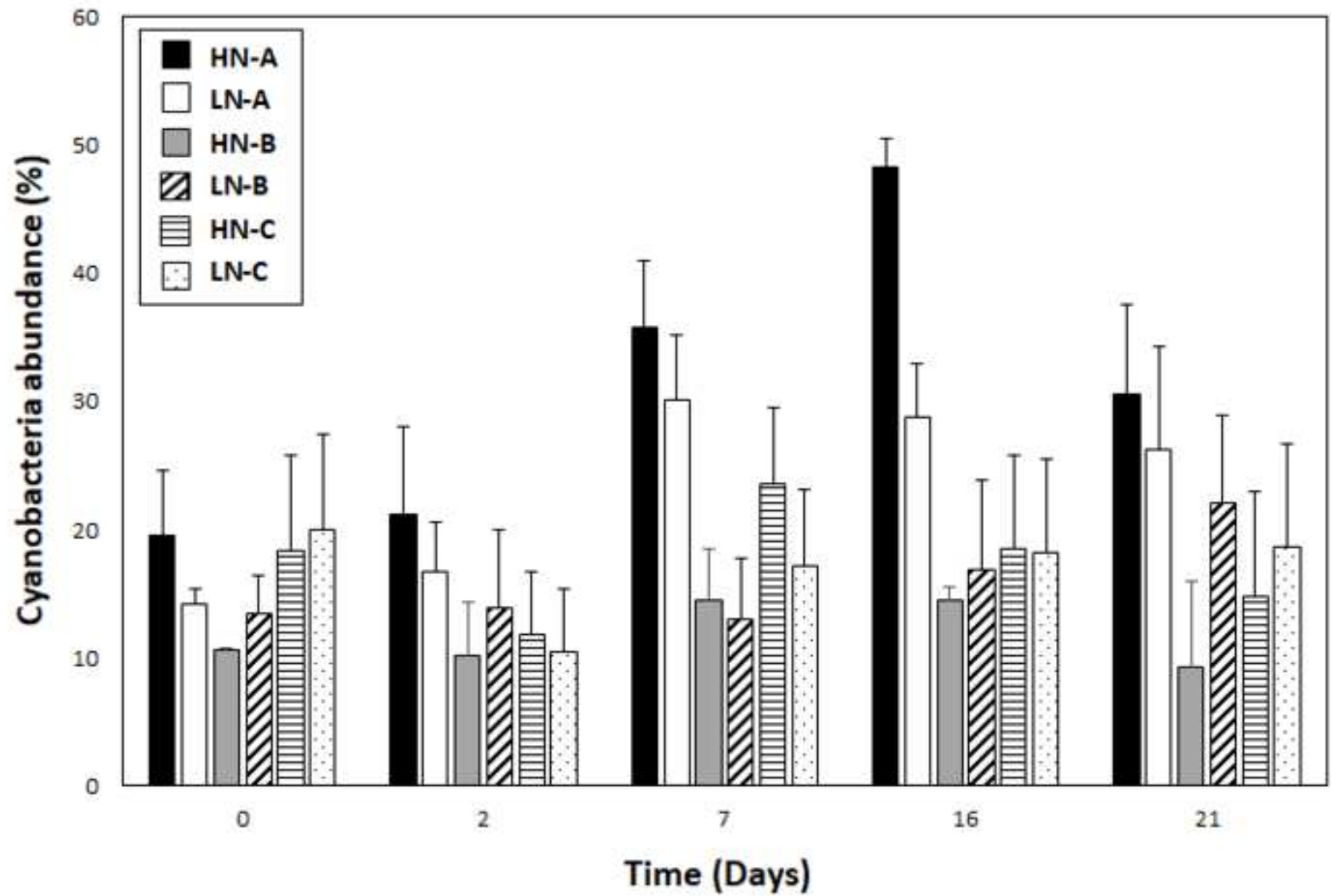
736 concentration in biofilm (ng/mg).

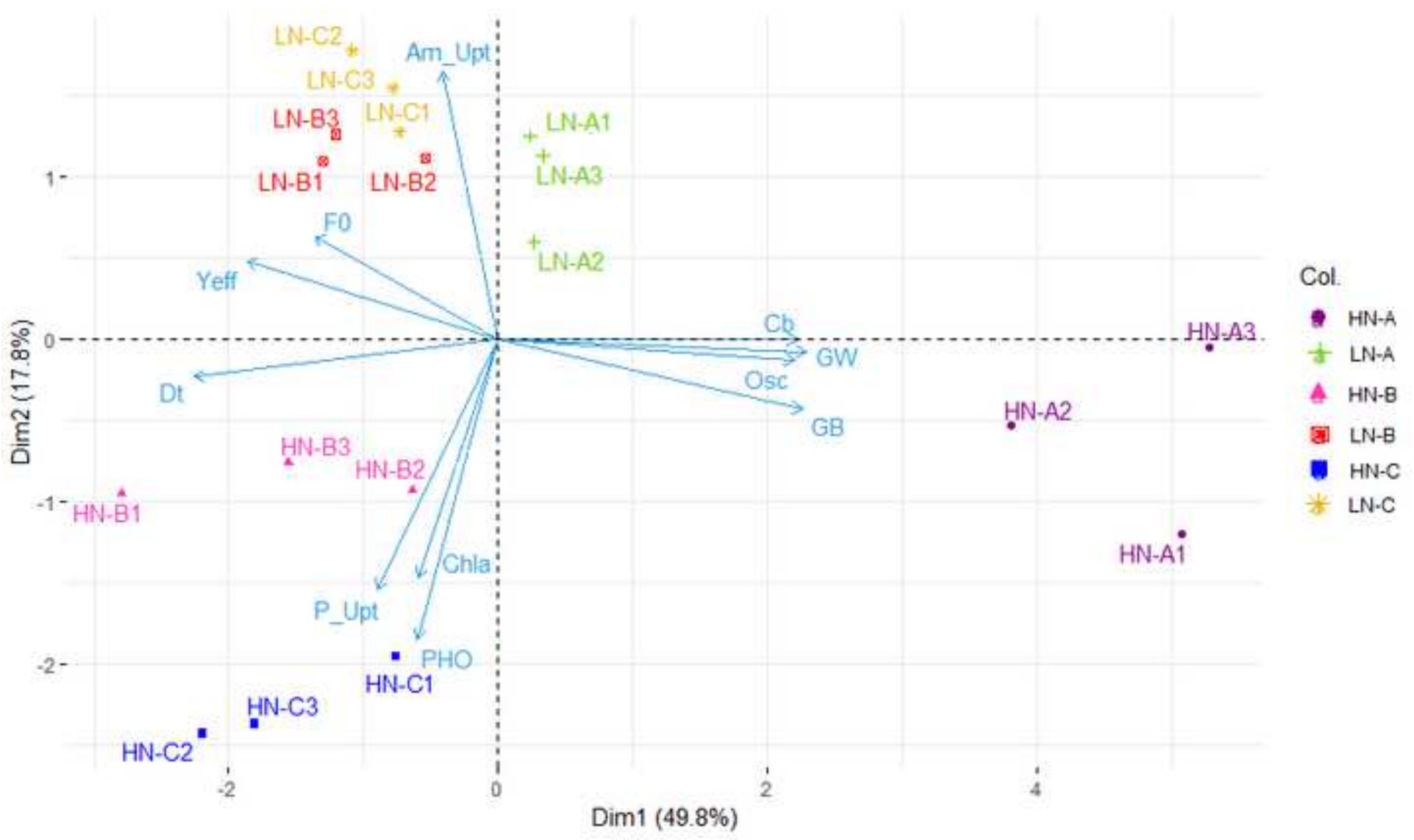
	Treatment					
	LN			HN		
	A	B	C	A	B	C
N-NH ₄ ⁺ (µg/L)	12.7 ± 2.7	12.6 ± 0.6	11.7 ± 1.7	110 ± 7	110 ± 6	106 ± 5
N-NO ₃ ⁻ (µg/L)	107 ± 15	108 ± 18	112 ± 5	1175 ± 112	1174 ± 90	1142 ± 24
P-PO ₄ ³⁻ (µg/L)	61.0 ± 4.8	18.2 ± 3.8	4.0 ± 0.6	682 ± 49	164 ± 9	42.0 ± 6.6
N:P	4.2 ± 0.2	14.9 ± 1.4	69.8 ± 7.2	4.4 ± 0.5	16.5 ± 1.1	64.8 ± 5.2
pH	8.1 ± 0.2	8.1 ± 0.2	8.1 ± 0.2	8.0 ± 0.3	8.0 ± 0.2	8.1 ± 0.2
Temperature (°C)	16.8 ± 0.4	16.9 ± 0.3	16.8 ± 0.3	16.7 ± 0.3	16.8 ± 0.3	16.9 ± 0.3
EC (µS/cm)	164 ± 17	163 ± 14	158 ± 33	166 ± 13	164 ± 14	170 ± 12
OD (mg/L)	8.3 ± 0.5	8.0 ± 0.6	8.2 ± 0.5	8.2 ± 0.5	8.4 ± 0.5	8.0 ± 0.4
Saturation (%)	85 ± 5	84 ± 7	85 ± 5	84 ± 5	86 ± 5	83 ± 4

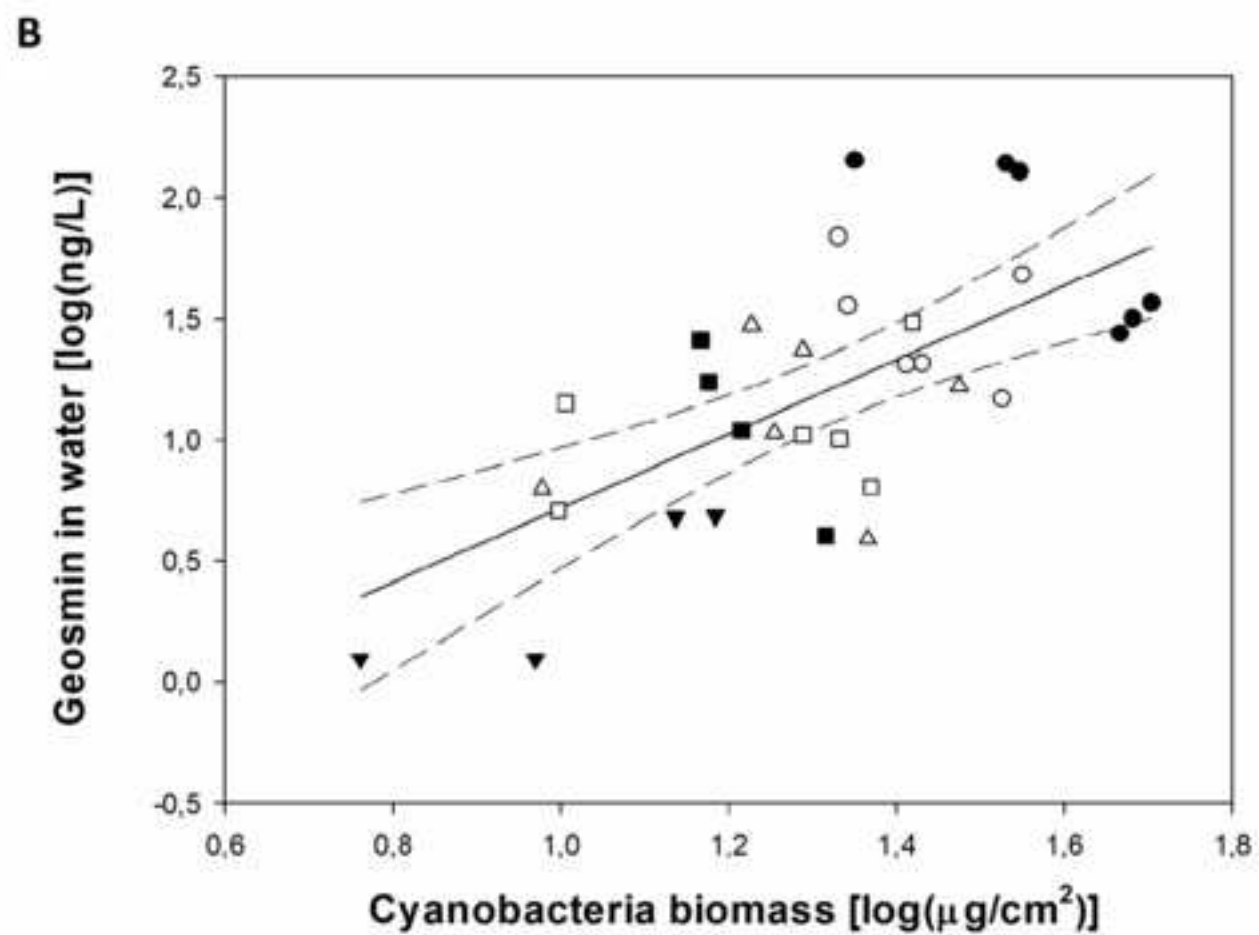
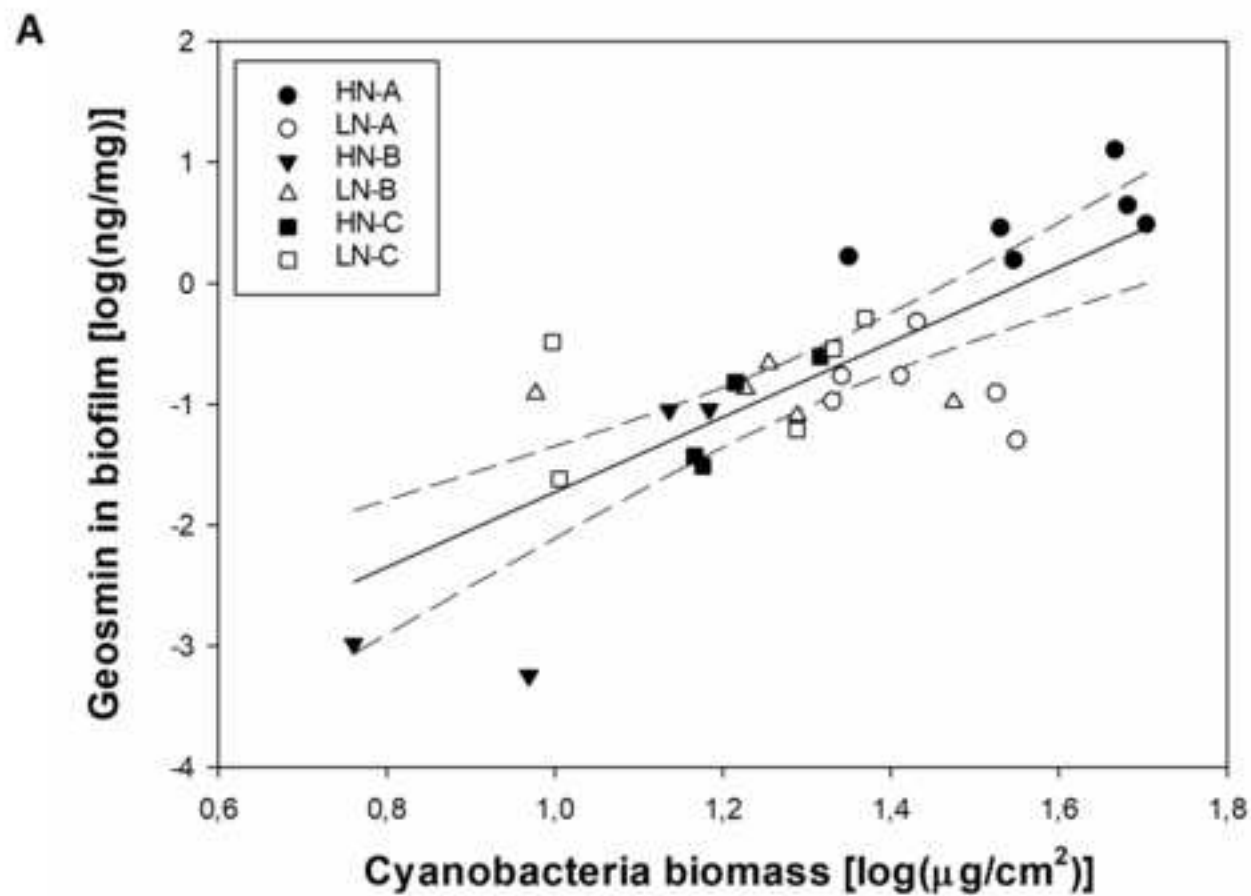
	Treatment																		
	LN									HN									
	A			B			C			A			B			C			
	0d	7d	16d	21d	7d	16d	21d	7d	16d	21d	7d	16d	21d	7d	16d	21d	7d	16d	21d
Geosmin in water (ng/L)	27.4 ± 6.7	21.2 ± 2.9	18.7 ± 2.8	20.9 ± 13.6	12.8 ± 3.4	7.0 ± 2.8	23.3 ± 5.4	14.8 ± 2.9	7.2 ± 2.1	18.4 ± 8.7	26.6 ± 2.2	32.0 ± 3.9	135.8 ± 6.1	7.5 ± 3.1	4.8 ± 0.0	< 2.5	11.6 ± 1.9	7.5 ± 3.5	21.5 ± 4.2
Geosmin in biofilm (ng/mg)	0.10 ± 0.05	0.03 ± 0.01	0.26 ± 0.19	0.08 ± 0.04	0.16 ± 0.08	0.17 ± 0.07	0.10 ± 0.03	0.13 ± 0.1	0.38 ± 0.12	0.04 ± 0.03	1.39 ± 0.35	3.76 ± 0.96	2.03 ± 0.74	0.08 ± 0.00	0.09 ± 0.00	0.00 ± 0.00	0.04 ± 0.03	0.20 ± 0.07	0.03 ± 0.00
Cyanobacteria (µg chla/cm²)	0.43 ± 0.06	1.23 ± 0.71	0.61 ± 0.27	0.55 ± 0.06	0.28 ± 0.22	0.38 ± 0.21	0.48 ± 0.20	0.52 ± 0.30	0.52 ± 0.10	0.39 ± 0.20	1.95 ± 0.36	1.63 ± 0.36	0.87 ± 0.38	0.28 ± 0.16	0.57 ± 0.15	0.46 ± 0.05	0.66 ± 0.23	0.34 ± 0.31	0.59 ± 0.14
Diatoms (µg chla/cm²)	2.10 ± 0.60	2.68 ± 0.42	1.55 ± 0.82	1.61 ± 0.45	2.37 ± 0.99	1.62 ± 0.63	1.63 ± 0.35	3.00 ± 2.31	2.66 ± 1.29	1.61 ± 0.51	3.49 ± 0.40	2.14 ± 1.12	2.33 ± 1.90	1.65 ± 1.02	4.35 ± 1.05	2.99 ± 2.52	2.00 ± 0.51	1.66 ± 1.69	3.38 ± 0.75
Chlorophyll a (µg/cm²)	3.90 ± 0.64	3.50 ± 1.31	3.10 ± 1.89	6.51 ± 0.51	2.63 ± 0.48	3.82 ± 1.19	4.20 ± 0.92	7.87 ± 0.88	4.14 ± 1.38	2.80 ± 0.70	2.87 ± 0.81	3.68 ± 0.32	8.68 ± 1.73	5.12 ± 2.45	7.99 ± 1.32	18.32 ± 0.97	5.10 ± 0.28	11.72 ± 2.31	10.00 ± 3.37
Yeff	0.589 ± 0.029	0.646 ± 0.064	0.581 ± 0.083	0.640 ± 0.053	0.614 ± 0.036	0.635 ± 0.068	0.680 ± 0.040	0.639 ± 0.047	0.555 ± 0.056	0.645 ± 0.033	0.630 ± 0.084	0.545 ± 0.042	0.529 ± 0.046	0.607 ± 0.083	0.657 ± 0.025	0.672 ± 0.031	0.633 ± 0.031	0.615 ± 0.036	0.631 ± 0.068
PHO (MUF/cm²-h)	19.7 ± 1.9	16.8 ± 7.7	16.5 ± 9.4	24.2 ± 4.2	13.0 ± 1.6	23.6 ± 12.9	29.3 ± 11.6	33.2 ± 8.5	42.0 ± 20.1	31.6 ± 11.7	17.4 ± 9.6	25.6 ± 9.2	35.9 ± 11.7	24.2 ± 8.3	17.5 ± 7.9	36.1 ± 6.5	26.9 ± 14.2	54.4 ± 28.8	72.0 ± 18.4

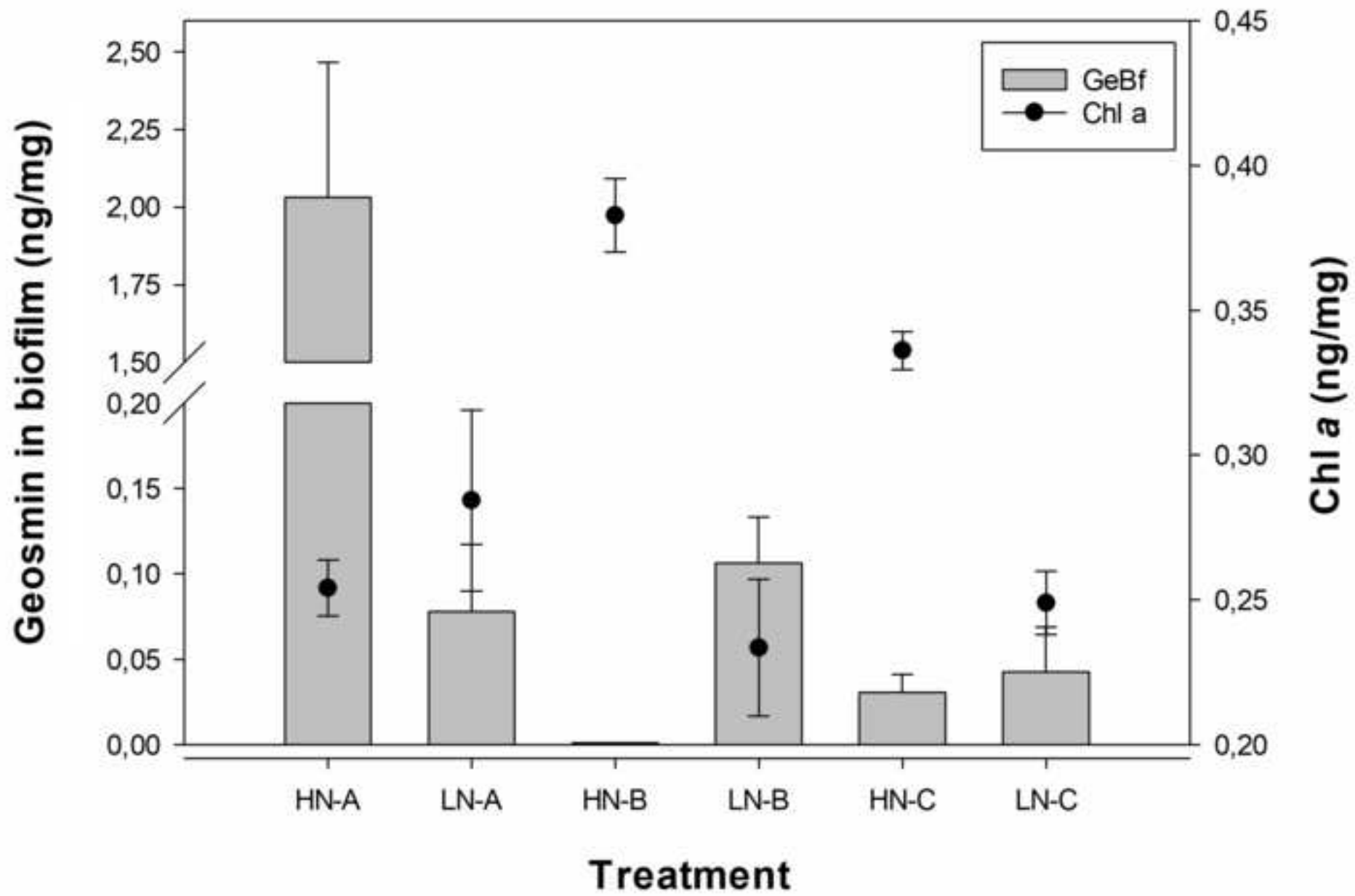
		Repeated measures ANOVA		ANOVA					
				t = 7d		t = 16d		t = 21d	
		F	p	F	p	F	p	F	p
Geosmin in Water (ng/L)	N	9.22	<0.001	0.34	n.s.	4.22	n.s.	21.65	0.001
	R	37.03	<0.001	22.54	<0.001	50.42	<0.001	126.28	<0.001
	NxR	15.30	<0.001	3.44	n.s.	7.47	0.01	50.05	<0.001
Geosmin in Biofilm (ng/mg)	N	26.34	<0.001	22.40	0.001	30.63	0.001	6.90	<0.05
	R	22.00;	<0.001	25.14	<0.001	39.07	<0.001	9.64	0.01
	NxR	24.89	<0.001	35.33	<0.001	39.95	<0.001	9.67	0.01
Cyanobacteria ($\mu\text{g chl}a/\text{cm}^2$)	N	2.80	<0.05	0.81	n.s.	5.08	<0.05	0.12	n.s.
	R	6.45	<0.001	5.28	<0.05	33.67	<0.001	3.44	n.s.
	NxR	1.53	n.s.	0.09	n.s.	7.75	<0.01	0.55	n.s.
Diatoms ($\mu\text{g chl}a/\text{cm}^2$)	N	0.98	n.s.	0.93	n.s.	1.75	n.s.	0.73	n.s.
	R	3.16	<0.001	4.39	<0.05	28.45	<0.001	3.44	n.s.
	NxR	1.76	n.s.	0.13	n.s.	12.00	<0.01	1.18	n.s.
Chlorophyll <i>a</i> ($\mu\text{g}/\text{cm}^2$)	N	41.04	<0.001	0.28	n.s.	14.82	<0.01	110.16	<0.001
	R	13.74	<0.001	12.08	<0.01	6.34	<0.05	14.53	0.001
	NxR	13.74	<0.001	6.57	<0.05	3.74	n.s.	22.19	<0.001
Yeff	N	3.72	0.01	0.78	n.s.	4.59	n.s.	2.06	n.s.
	R	4.86	<0.001	0.30	n.s.	6.00	<0.05	5.24	<0.05
	NxR	2.23	<0.05	0.02	n.s.	0.99	n.s.	3.16	n.s.
PHO ($\mu\text{mol MUF}/\text{cm}^2\cdot\text{h}$)	N	1.82	n.s.	0.17	n.s.	0.38	n.s.	11.17	<0.01
	R	0.73	n.s.	3.48	n.s.	4.96	<0.05	5.77	<0.05
	NxR	1.14	n.s.	1.22	n.s.	0.43	n.s.	3.29	n.s.













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CE: Conceptualization, Methodology, Writing-Original Draft Preparation, **MA:**

Conceptualization, Investigation, **SP:** Supervision, **MR:** Investigation, Resources, **LV-P:**

Investigation, Methodology, **MO:** Supervision, **LL:** Supervision, **LP:** Conceptualization,

Investigation, Writing.

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: